



Graphene oxide–polysulfone filters for tap water purification, obtained by fast microwave oven treatment

Downloaded from: <https://research.chalmers.se>, 2025-12-04 14:41 UTC

Citation for the original published paper (version of record):

Kovtun, A., Zambianchi, M., Bettini, C. et al (2019). Graphene oxide–polysulfone filters for tap water purification, obtained by fast microwave oven treatment. *Nanoscale*, 11(11): 22780-22787. <http://dx.doi.org/10.1039/c9nr06897j>

N.B. When citing this work, cite the original published paper.

Graphene oxide-polysulfone filters for tap water purification, obtained by fast microwave oven treatment.

Alessandro Kovtun,^{a†} Massimo Zambianchi,^{a†} Cristian Bettini,^a Andrea Liscio,^b Massimo Gazzano,^a, Franco Corticelli,^b Emanuele Treossi,^a Maria Luisa Navacchia,^a Vincenzo Palermo,^{a,c,*} Manuela Melucci^{a,*}

The availability of clean, pure water is a major challenge for the future of our society. 2-dimensional nanosheets of GO seem promising as nanoporous adsorbent or filters for water purification; however, their processing in macroscopic filters is challenging, and their cost vs. standard polymer filters is too high. Here we describe a novel approach to combine graphene oxide (GO) sheets with commercial polysulfone (PSPSU) granules for improved removal of organic contaminants from water. The adsorption physicphysics of contaminants on the PSPSU-GO composite follows the Langmuir and Brunauer-Emmet-Teller (BET) modelmodels, with partial swelling and intercalation of the molecules in between GO layers. Such mechanism, well-known in layered clays, has not been reported previously for graphene or GO. Our approach requires minimal amounts of GO, deposited directly on the surface of the polymer, followed by stabilization using microwaves or heat. The purification efficiency of the PSU-GO-PS composites is significantly improved vs. benchmark commercial PSPSU, as demonstrated by removal of two model contaminants, Rhodamine B and Ofloxacin. The excellent stability of the composite is confirmed by extensive (100 hours) filtration tests (100 h) in commercial water cartridges.

The outstanding adsorption efficiency of graphene oxide (GO) toward organic molecules has been widely proved in the last few years.¹⁻⁷ The high surface area and the multiple surface chemical groups enable strong interface interactions by π - π stacking through the sp^2 region, by H bond through the carboxyl, hydroxyl or epoxy groups as well as van der Waals and hydrophobic interactions.^{3, 8, 9} Such unique features have motivated the extensive use of graphene (both GO and reduced GO, rGO) and polymer-GO/rGO 3D composites such as hydrogel/aerogel,^{10, 11} sponges,¹²⁻¹⁴ membranes¹⁵⁻²⁰ as adsorbent for water remediation. The plethora of reported data is highlighting highlights graphene and GO as the most promising nanomaterials for the development of advanced water treatment technologies, both drinking and wastewater, technologies.²¹⁻²³ Key advantages of hybrid GO based 3D structures are: i) the easy recovering of the adsorbent after use, ii) the significant enhancement of adsorption performance by addition of GO at even at low amount, iii) the possible modification of GO for promoting selective binding processes.²⁴⁻³⁰

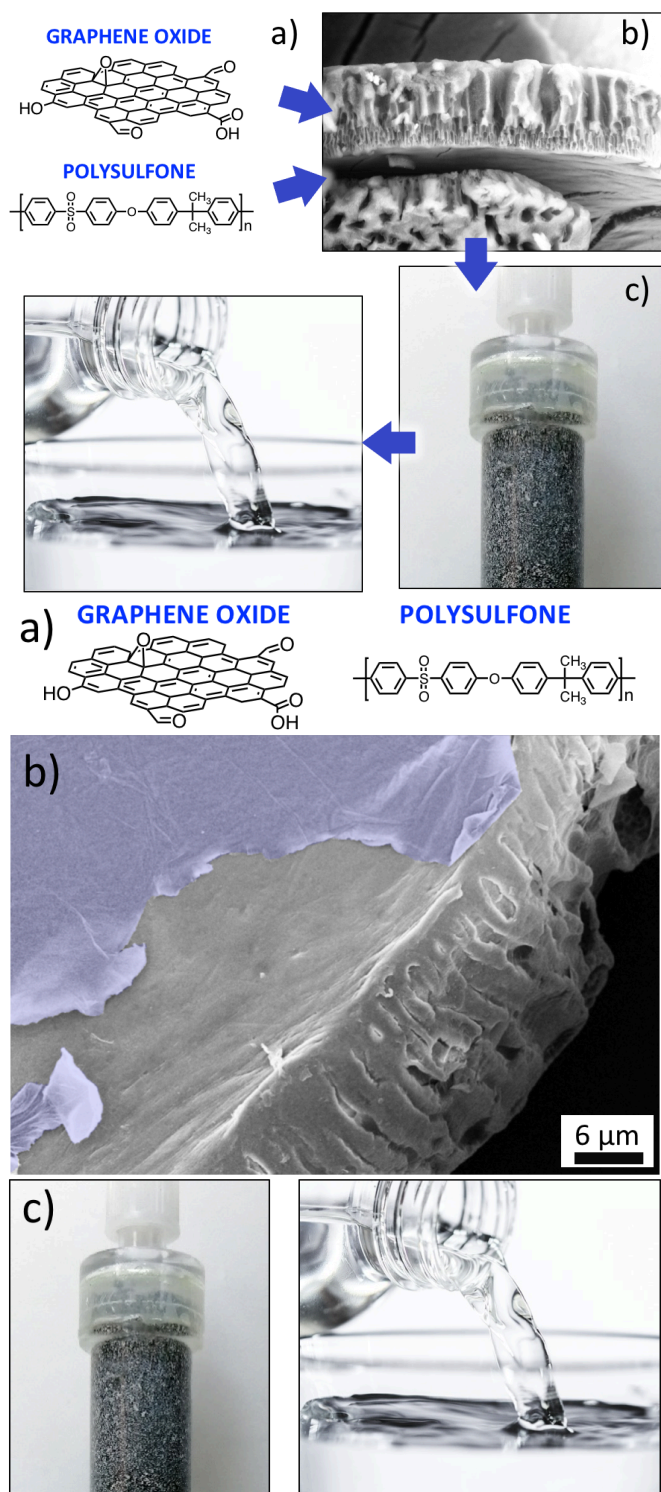


Figure 1. a) Representative chemical structure of GO and PPSU. b) SEM and c) optical image of the PPSU-GO composite filter. The GO sheets have been highlighted in blue in b). SEM unfiltered images are available in SI.

We recently demonstrated the superior performance of polysulfone-GO (PPSU-GO) composites for the removal of several classes of emerging organic contaminants from tap water (figure 1).^{31, 32} These compounds are currently cause of major environmental concern due to their increasing occurrence in surface water bodies and even drinking water and to their potential effect on human health and ecosystem.³³⁻³⁵ The composite was obtained by phase inversion of a PPSU-GO mixture (5%w/w of GO) using water and N-methyl-2-pyrrolidinone (NMP) as orthogonal solvents, i.e. the standard industrial

method for the preparation of PSU ultrafiltration membranes. The contaminants captured could then be removed by simple washing of the composite in ethanol,³¹ allowing in this way to recycle/reuse the filter material.

A major challenge in developing new materials for filters for water filter remediation is to ensure their safety, i.e. utilize cheap and environment friendly methods to produce them, and stabilize the material so that no poisonous debris or molecules are released, even after extensive use.

Building on our previous results, here we describe a new method to stabilize GO nanosheets on PS using microwaves or heat. The procedure is entirely performed in water. The method is based on directly coating GO on the surface of cheap industrial scraps of PS. PSU hollow fibers with GO³² Instead of using phase inversion, we fix the GO nanosheets on the polymer by use of microwaves or thermal/conventional heat treatments. The procedure is entirely performed in water.

Such approach allows to produce grams scale of PPSU-GO composite powder/material with no need of toxic solvents, on PSU filters already commercialized. We then tested the ability of such composites to capture two important organic contaminants (a textile dye and a medical drug) comparing them in comparison with standard, commercial PS/unmodified PSU granules, unravelling this allowing to unravel the physics of the adsorption process. We finally tested the stability of the composite upon prolonged water flow operation, by extensive test in water purification filters.

Coating of GO on PPSU recycled granules

We used as starting substrate the scraps of the industrial production of polysulfone ultrafiltration hollow fibers ultrafiltration membranes; made of PSU (representing about 10% of the total yearly production); this substrate, even if more challenging and complex than polymeric substrates we used in previous works,^{36, 37} has the key advantage, as industrial waste, of having negative costs and represents an environmental added value of the final composite.³² Polysulfone hollow fiber (MediSulfone®) scraps were mechanically ground to obtain polysulfone hollow porous granules (figure S1a-c, ESI). Polysulfone hollow fiber scraps (MediSulfone) were mechanically ground to obtain polysulfone hollow porous granules (figure S1c).

The granules were then dispersed in a solution of GO in DI water; GO concentration was tuned to have 5 wt % GO content in the final composite. Water was removed by heating at 50 °C on a rotary evaporator.

The granules were then dispersed in a solution of GO in DI water, prepared as described in ESI; GO concentration was tuned to have 5 wt % GO content in the final composite in order to allow direct comparison with what already described in previous work.³¹ Water was removed by heating at 50 °C on a rotary evaporator.

Stabilization of GO coatings on PPSU

A main challenge in graphene-polymer composites is to ensure a good interaction between the nanosheets and the polymer, typically to enhance mechanical properties.³⁸ In composites for water filters, a good interaction is required to ensure a long-lasting adhesion of the nano-additives to the matrix, to avoid their release in the treated water. Graphene and GO are excellent adsorbers of microwaves, and we exploited this opportunity in the past to achieve fast, environmentally free functionalization of GO.³⁹ Here, we do not use microwaves for a specific functionalization of GO in solution, but to ensure fixation of GO nanosheets to the PSU scraps, thus ensuring a good cohesion of performance and stabilizing the composite.

The PS stabilization was performed on the PSU-GO powder was irradiated with microwaves at atmospheric pressure in a CEM Discover SP apparatus ($f = 2.45$ GHz) which has in situ magnetic variable

speed, irradiation monitored by PC computer, infrared temperature measurement and continuous feedback temperature control. Samples were irradiated for 45 min at 100 W (fixed power).. During MW irradiation the temperature was measured, and never exceeded 70 °C. The PPSU-GO was finally washed in a 1:1 mixture of water/ethanol to remove any traces of unreacted GO and finally left to dry at room temperature until weight stabilized. (PSU-GO-MW, Figure 1).

In an alternative approach, PPSU-GO was also treated in a standard oven for 2 hours at 120 °C. The material was allowed to cool to room temperature, then washed and dried as described above. (figure S1d, ESI).

Remarkably, the The presence of GO coating was stable only after the MW or oven annealing treatment confirmed by high resolution Scanning Electron Microscope (SEM) images (figure S2), ESI). No changes in surface porosity and cross-section were observed in comparison to pristine PS granules after coating, as measured by standard techniques (N₂ adsorption).

Remarkably, both treatments enhanced greatly the stability of GO on PSU, as shown by extended immersion in water (figure S3, ESI).

The pristine PPSU-GO composite, the composite treated with microwaves and the one treated in oven (named respectively Blank, PSGO PSU, PSU-GO-MW and PSGOPSU-GO-OV) were then characterized by XPSXRD and XRD XPS.

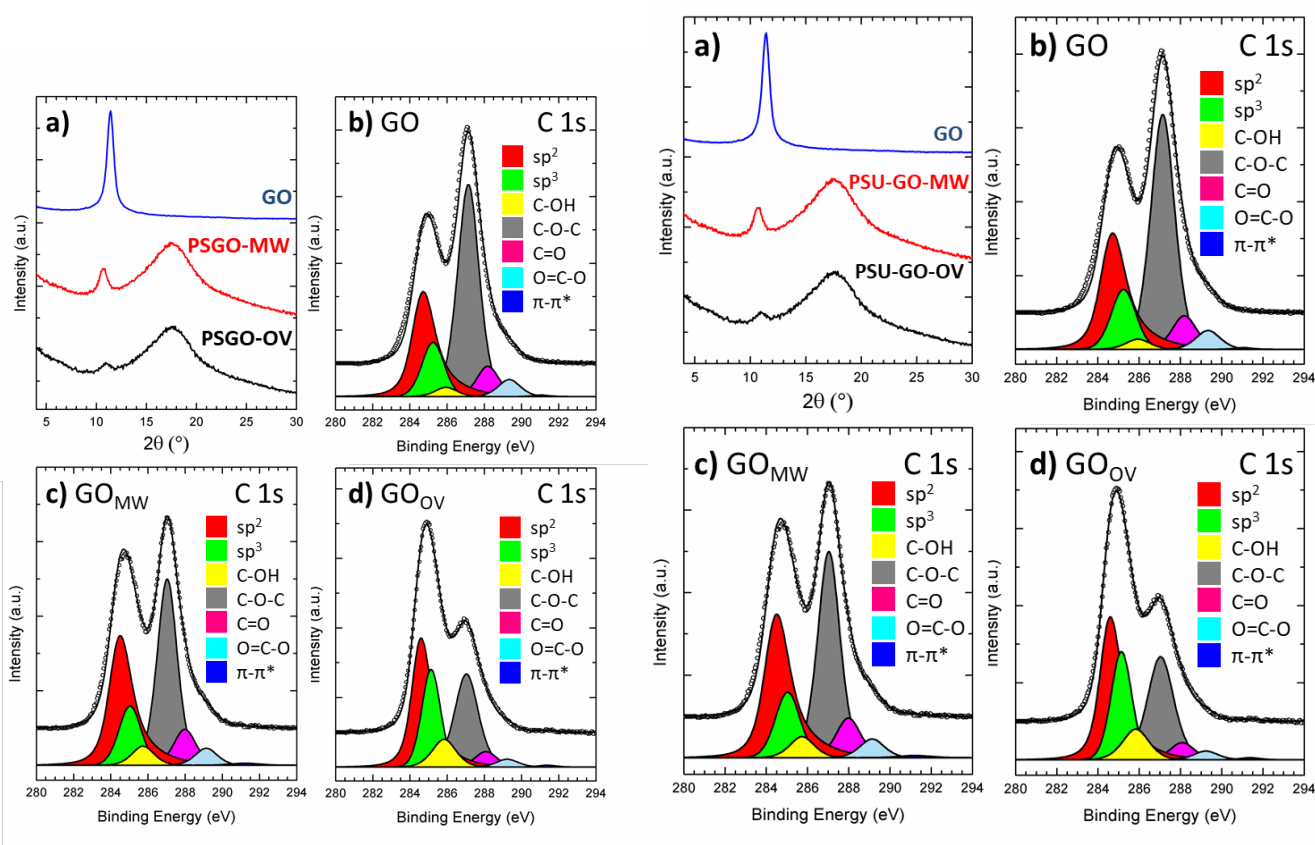


Figure 2. (a). XRD pattern of GO (blue line), PSGOPSU-GO-MW (red line) and PSGOPSU-GO-OV (black line). The (0 0 1) peak of the GO structure is well appreciated in both PSGOPSU-GO samples, at 10-12°, together with the amorphous bell-shaped band due to PS.PSU at 15-20°. XPS C 1s spectra of b) pristine GO, c) PSGOPSU-GO-MW samples after polymer dissolution with DCM and d) PSGOPSU-GO-OV samples after polymer dissolution with DCM.

Structural characterization of the PSGOPSU-GO composites

X-ray diffraction patterns were collected in Bragg-Brentano geometry (CuK radiation, 0.15418 nm). Average thickness of GO crystal stacks was calculated from peak width by using the Scherrer equation. The number of GO layers was estimated as the ratio of the size and the interlayer distance, obtained from position of (001) GO peak.

Figure 2a compares the XRD of pure GO with the PSGOPSU-GO-MW and PSGOPSU-GO-OV composites. Pure GO showed a sharp, intense peak, as expected for a bulk 2D material. An average spacing of 7.8 ± 0.3 Å was observed, with average of 14 ± 2 GO layers stacked on top of each other.

PSGOPSU-GO-MW and PSGOPSU-GO-OV samples showed the amorphous bell-shaped band due to PPSU, but also a clear (001) peak of the GO structure, indicating the presence of stacked GO multilayers on the surface of the PPSU granules. The stacking distance was 8.3 ± 0.3 Å for PSGOPSU-GO-MW and 7.9 ± 0.3 Å for PSGOPSU-GO-OV; the estimated number of layers was ca. 10 layers, slightly lower than what observed in blank sample (see Table S2S1, ESI for full details and exp. errors).

The fast microwave or oven treatments did not change the mesoscale structure of the membranes, with no deformation or destruction of the PSU microchannels as observed by SEM. No changes in surface porosity and cross-section were observed in comparison to pristine PSU granules after coating, as measured by standard techniques (N_2 adsorption).

The differences between the two methods were instead observed clearly by XPS.

Chemical characterization of the PSGOPSU-GO composites

XPS is a powerful technique, able to probe the outer surface of the granules, thus allowing to get information on the possible presence of the GO coating, its uniformity, its oxygen/carbon ratio and the abundance of different chemical groups. We performed XPS on all the samples, and analysed the results with a novel deconvolution procedure which allow to obtain precise chemical analysis and oxygen content in GO also in presence of other oxygen-containing materials (figureFigure 2b,c,d and Table 1S2).⁴⁰ Such procedure is based on quantitative line-shape analysis of C 1s signals with asymmetric pseudo-Voigt line-shapes (APV), in contrast to Gaussian-based approaches conventionally used (see ESI and ref.⁴⁰ for details).

The XPS of untreated, bulk PPSU granules yielded carbon/oxygen and sulphur/carbon O/C= 0.13 ± 0.01 and S/C= 0.04 ± 0.01 in good agreement with PPSU chemical composition (O/C=0.15 and S/C=0.04, respectively).

XPS of pristine GO (Figure 2b) showed the presence of sp^2 areas and several chemical groups typical of GO, with an average oxygen/carbon ratio O/C= 0.39 ± 0.01 . (see also figs. S4,5 and table S2, ESI).

XPS technique probes only the outer surface (2-3 nm) of the powders; thus, from the contribution to XPS data of GO and PPSU we could estimate the amount of surface composed of uncoated PPSU and the one coated from the GO nanosheets. The GO coating covered >50% of the surface of the granules; however, it was not possible to obtain a more precise estimateestimation.

To overcome this issue and have a detailed analysis of the GO coating, we thus had to remove completely the polymer phase. This was achieved by immersion in dichloromethane (DCM). This); PSU is soluble in DCM while GO is stable in such solvent completely removes , allowing a more precise characterization of the PS, while leaving unaltered the oxidation state of GO. remaining GO by XPS. For comparison sake, pure GO was also immersed in DCM and analyzedanalysed by XPS. Table 1S2 shows the O/C ratio measured for all chemical composition of the different samples, including their O/C ratio. We see that, while treatment with MW does not alter significantly the GO chemistry (O/C= 0.38 ± 0.01),

classical heating in oven, even if at a relatively low temperature of 120°C, gives partial reduction of the GO, with the O/C ratio decreasing to 0.30±0.01.

Table 1: Oxygen/Carbon ratio measured by fitting the XPS C 1s peak on different samples. Exp. error is ±0.01 in all samples.

Material	O/C ratio
Blank PS *	0.13
GO - pristine	0.39
GO washed in DCM	0.40
GO _{MW} <i>after PS dissolution in DCM</i>	0.38
GO _{OV} <i>after PS dissolution in DCM</i>	0.30

* O/C of blank PS was obtained from the area ratio of O 1s and C 1s due to the significant presence of S-O groups in PS structure.

Contaminants adsorption tests

The adsorption capacities of blank PS, PSGOPSU, PSU-GO-MW and PSGOPSU-GO-OV samples were studied by exposing them to tap water contaminated with different emerging organic contaminants (EOC). The removal performance of PSU-GO towards many widely used contaminants (i.e. drugs, pigments, surfactants etc.) was previously described in ref.³¹; here we focused more on the preparation technique of the composite and on their application in water cartridges. As test contaminants, we chose Rhodamine B (RhB, a dye largely used in textile and pharmaceutical industries), and ofloxacin (OFLOX, a quinolonic antibiotic). Structure of both molecules is shown in Figure 3.

Both molecules are of significant concern for pollution of surface water bodies.,⁴¹ Structure of both molecules is shown in figure 3. and adsorption capacity of OFLOX and RhB has been already reported for GO sheets.^{42,43}

Using 100% GO in the cartridges to remove such molecules would be effective, but bulk GO in powder cannot be extruded or processed in pellets or fibers, it can be volatile and also burn or explode in suitable conditions. High adsorption capacity of OFLOX and RhB was the filtration cartridge and problems of material aggregation or mechanical stability would also hinder straightforward application of GO, making it incompatible with actual cartridge production technique, which is based on PSU fibers. Conversely, the advantage of our method is that can be applied not only on commercial membranes already reported for GO sheets.^{42,43} We thus produced on large scale, but also on scraps deriving from their preparation, as demonstrated in the previous sections.

We tested the adsorption performance of pure GO, blank PS, PSGOPSU, PSU-GO-MW and PSGOPSU-GO-OV composites by dedicated experiments performed in tap water contaminated on purpose spiked with the target molecules (Figs. S10, ESI).. Experimental data were then fitted using Brunauer–Emmett–Teller (BET, eq. 1), Langmuir (eq. 2) and Freundlich models, see tables S5-S10. Details of HPLC method used for analytical determination are reported in ESI.

The adsorption isotherm of GO was performed at fixed concentration of RhB and OFLOX by varying the amount of GO. The two contaminants (in powder, as received) were added to GO suspensions (5 ml) at different concentrations. The details of sample preparation and a list of samples prepared are reported in tables S3 and S4. The isotherm curves were obtained equilibrating the solutions for 24 hrs at room temperature. The solutions were analysed by HPLC.

The adsorption isotherms of blank PS, PSGO-MW and PSGO-OV were fitted with the Brunauer–Emmett–Teller (BET) adsorption model (eq. 1, figure 3). The BET model describes a multilayer adsorption mechanism in the gas–solid and liquid-solid equilibrium systems. described by the formula: ⁴⁴

$$Q_e = (Q_m \cdot C_{BET} \cdot x) / ((1-x) \cdot (1 + C_{BET} \cdot x)) \quad (1)$$

Here Q_e (mg/g) is the quantity amount of molecules adsorbed when the equilibrium concentration is C_e (mg/mL); Q_m is the, the quantity amount of molecules needed to cover the entire adsorbent surface with a monolayer (i.e. the monolayer saturation capacity, mg/g). The thermodynamic equilibrium BET constant is $C_{BET} = \exp(\Delta E / K_b T)$, where ΔE is the difference between the energy of adsorption of first and second layers of molecules, $K_b T$ is the thermal energy; $x = C_e / C_s$, where C_s (mg/mL) is the adsorbate saturation concentration in BET adsorption process (see ref. ⁴⁵). The linear fit was performed plotting $y = 1 / (Q_e \cdot (1/x - 1))$ vs. x and the results are shown in figure 3 and in table S3x.

Langmuir model describes a monolayer adsorption process, which depends on the equilibrium concentration C_e (mg/mL); the Q_m (mg/g) and the thermodynamic equilibrium constant K_L (mL/mg).

$$Q_e = Q_m \cdot K_L \cdot C_e / (1 + K_L \cdot C_e) \quad (2)$$

Remarkably, Eq. 1 can fit very well all adsorption isotherms tested with of Ofloxacin on the 3 three different materials and 2 target molecules.. This indicates that the adsorption follows in all these cases the BET physical model, i.e. can form multilayers on the substrate, filling small pores first.

Once assessing that On the same physical other hand, the adsorption of RhB on the 3 different materials can be better described by Langmuir model applies (eq. 2). It should be noted anyhow that both molecules show a good fit to both BET and Langmuir models (tables S8-S10, ESI). Freundlich isotherm model gave instead a poor fit for all the surfaces under test, we compared their performance in capturing RhB and OFLOX in solution. samples.

Figure 3 shows that both PSGOPSU-GO composites outperform the standard PS polymer PSU material. In particular, PSGOPSU-GO-MW samples captured ca. 3 times more RhB than blank, and >ca. 2 times more OFLOX than PSU.

We also compared It is interesting to compare the adsorption of RhB and OFLOX in solution with adsorption of N_2 molecules from gas, which is the conventional technique used to measure the specific surface area (SSA) of nanomaterials (table S11, ESI).

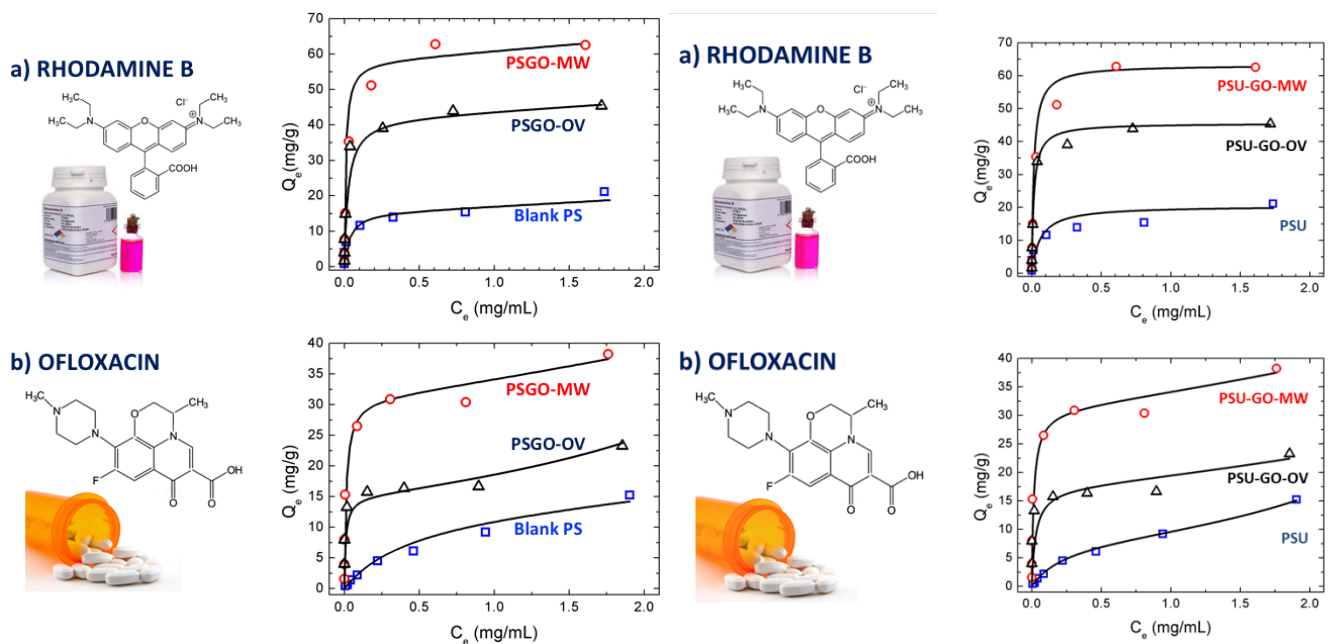


Figure 3. Chemical formulas/structures of RhB (a) and OFLOX (b), and the corresponding adsorption isotherms. Black lines represent Langmuir model for RhB (a) and BET model for OFLOX (b). The two processes are very different; N_2 adsorbs from gas phase, with weak interaction between N_2 molecules and the substrate, mainly due to Van der Waals interactions. In solution, instead, molecules adsorb by displacing other solvent molecules already present on the surface,⁴⁶ often interacting strongly with the substrate and the solvent due to electrostatic forces.

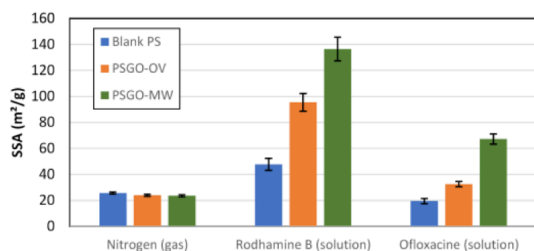
That's This explains why the surface area measured by gas adsorption of GO,⁴⁷ and more generally of layered materials,⁴⁸ is systematically smaller than what measured by adsorption of molecules in solution. To compare the adsorption capability of different molecules, we converted the quantity of molecules adsorbed Q_m (measured above in mg/g) in the SSA ideally occupied by those molecules using:

$$SSA = Q_m M_w^{-1} N_A A_{mol} \quad (23)$$

where M_w is the molecular weight, N_A is the Avogadro number and A_{mol} is the estimated "footprint" of the molecule, i.e. the amount of substrate ideally occupied by a single adsorbed molecule. We could estimate $A_{mol} \approx 1.81 \text{ nm}^2$ for RhB and $A_{mol} \approx 1.30 \text{ nm}^2$ for OFLOX, based on STM⁴⁹ and XRD measurements⁵⁰ respectively. This is just an approximation, and the actual footprint of the molecules on the surface will likely vary due to specific interactions between the molecule, the solvent and GO.⁵¹ We expect though that the footprint will remain of the same order of magnitude of what was previously reported.^{49 50}

Figure 4a compares the estimated SSA of blank PS, PSGOPSU, PSU-GO-MW and PSGOPSU-GO-OV, estimated by molecular adsorption of N_2 from gas, RhB and OFLOX from water solutions. We underline that the SSA reported should be considered as specific to the material studied, to be used only for direct comparison among samples, because it depends not only on the material, but also on its interaction with the molecule considered.

a) Area available to molecular adsorption



Area available to molecular adsorption

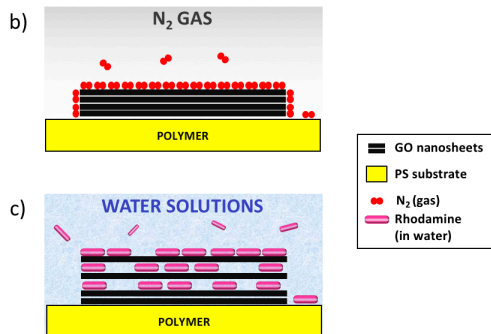
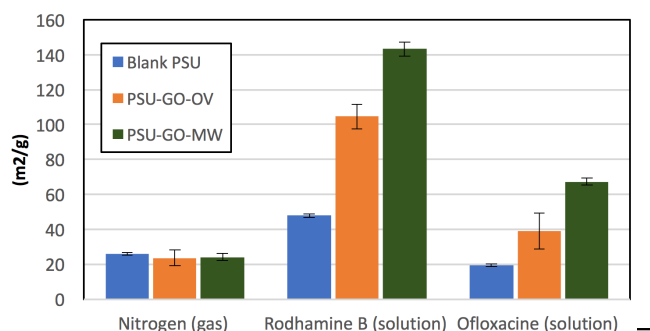


Figure 4. a) Comparison of the area measured in each specific molecule/substrate combination tested. Conversion from mg/g to m²/g has been performed estimating molecular footprint, as described in main text. b,c) Schematic cartoon showing the different possible capture mechanism of N₂ in gas and of organic molecules in solutions (see also ref. ⁴⁸)

In case of nitrogen adsorption from gas, all three materials give comparable SSA values, ≈ 23 - 25 m²/g, confirming that the GO coating and fixation does not change the overall porosity of the material, and that N₂ adsorption does not depend significantly on the surface chemistry of the different samples. OFLOX adsorption on pristine, blank PPSU gives SSA = 20 ± 2 m²/g, a value comparable to what measured with nitrogen; conversely, upon GO coating the number of molecules which can be captured by the material increases significantly. The estimated SSA increases to 3339 ± 2 m²/g for the oven-treated PSU-GO, and to 67 ± 4 m²/g for the MW treated PSU-GO, ca. three times the original value. RhB gives even more extreme changes, likely due to its electric charges (being a salt). The original SSA measured on blank (48 ± 5 m²/g) increases to 95105 ± 7 m²/g for oven treated PSU-GO, and to 135143 ± 9 m²/g for the microwave treated sample, corresponding to an adsorption capacity of 6063 mg of Rhodamine per gram of sample composite. Such SSA is more than five times larger than the SSA measured with nitrogen (24 m²/g). Such large difference cannot be explained uniquely by the uncertainty in the detailed molecular arrangement of the molecule on the substrates; furthermore, both RhB and OFLOX give significantly different SSA on different substrates, while adsorption of N₂ from gas gives similar values for PSGO, PSGO-MW and PSGO-OV.all samples.

Similarly, tests performed on 100% GO, with no PS, gave extremely high values of SSA, reported in table S6 (though if the use of pure GO would not be economically viable for filters).

The differences observed in figure 4 suggest that GO can act as an effective adsorbent for organic molecules thanks to its layered structure, able to capture the target pollutants for more than 6% of its weight in the best scenario (RhB on microwave treated samples). Adsorption test performed on pure GO also gave very high values of SSA (table S5).

It is known from characterization of layered minerals like clay that the presence of swelling minerals should be suspected when the SSA measurements in liquid are significantly higher than gas adsorption measurements.⁴⁸ GO is a unique layered material, which can easily be exfoliated in water;⁵² water molecules can intercalate and travel efficiently among stacked GO nanosheets.⁵³ Water trapped in between the GO nanosheets can behave as bulk water.⁵⁴

To demonstrate this hypothesis, we performed XRD scans on PSGOPSU-GO-OV composites exposed to a large excess of RhB contaminants (figure S9 in S8, ESI). In the exposed samples the (001) peak of GO almost disappeared, a clear indication that the organic molecules were disrupting the periodic stacking of the nanosheets. While intercalation of ions or molecules can give XRD peaks in graphite intercalation compounds, no periodicity could be observed in the PSGOPSU-GO samples exposed to RhB, likely due to the small number of nanosheets involved and the uneven intercalation of RhB within the GO layers.

GO stability, measured by release tests

Any material used for water purification should be safe, i.e. should not release additional contaminants in the outgoing water. A complete study of adsorbent materials should not only measure how much contaminants can transfer from the water to the filter; it should also demonstrate that there is no transfer of contaminants from the filter to the water. We observed in target experiments (figure S6S10, ESI) that our GO is “intrinsically safe”, while being highly soluble in distilled water, but is insoluble in tap water, thus being “intrinsically safe” for this specific application.

In addition to this evidence, we decided to test any possible release of GO by inserting the composite samples in a commercial water-filter cartridge⁵⁵ and flowing circulating water (both mQ and tap water) through the cartridge at 2L/h for 100 hours.

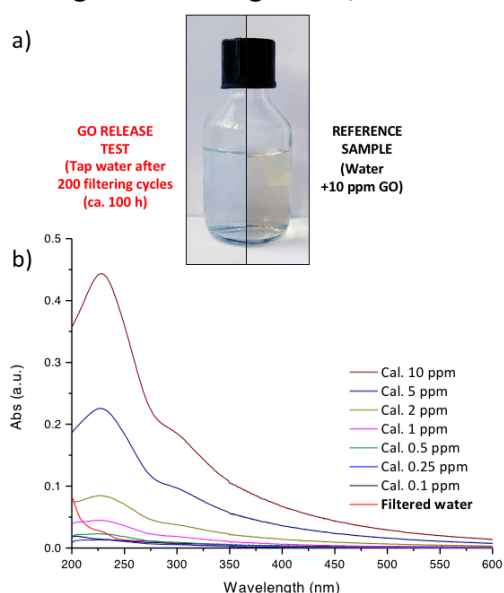


Figure 5. a) Photo of mQ water filtered for 100 h through the PSU-GO filter compared to standard solution of GO in mQ water at different concentration. b) UV-Vis absorption spectra of water filtered for 100 h through the PSU-GO filter, compared to calibration samples of water contaminated with different amounts of GO.

The filtered water was then analyzed to detect possible GO traces beyond such limit. For better safety assessment, we used two parallel GO detection techniques: UV-vis absorption and dynamic light scattering. We compared the filtered water to standard samples consisting of mQ water contaminated with known amounts of GO.

We performed UV-vis spectroscopy on water re-circulated for 100 hours through filters containing PSU-GO-MW and PSU-GO-OV powders (Figure S11, ESI). Detection limit, estimated with calibrated GO solutions, was about 1 ppm.

Figure 5 shows the images of the recirculated water, as well as the UV-vis spectra of the filtered water (red line) and of calibration solutions having a concentration range 0.25-10 ppm. The comparison indicates that any GO possibly released in filtered water was below 1 ppm.

Recent work indicate that safe limits of GO concentration to avoid toxic effects are between few tens and few hundreds of 10-1000 ppm (for a complete review on this important topic see ref. ⁵⁶. GO limits for aquatic organisms are in the range 40-2000 ppm.⁵⁷

Experiments in tap water were also performed showing transparent solution even after concentration of the sample (Figure S15, ESI).

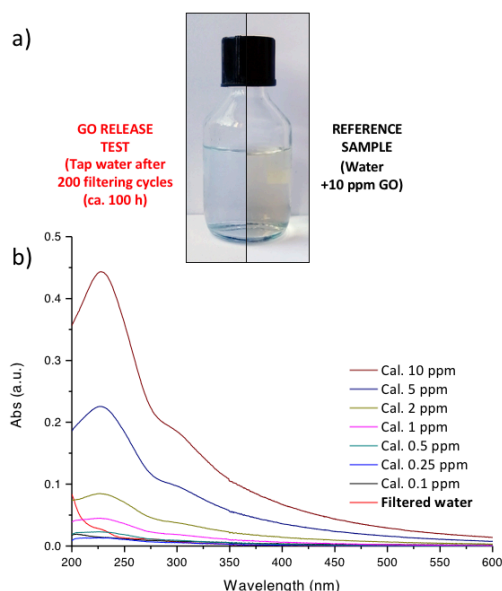


Figure 5. a) Photo of tap water filtered for 100 h through the PSGO filter compared to water contaminated with 10 ppm of GO. b) UV-Vis absorption spectra of water filtered for 100 h through the PSGO filter, compared to calibration samples of water contaminated with different amounts of GO.

The filtered water was then analyzed to detect possible GO traces beyond such limits. For better safety assessment, we used two parallel GO detection techniques: UV-vis absorption and dynamic light scattering. We compared the filtered water to standard samples consisting of mQ water contaminated with known amounts of GO.

We performed UV-vis spectroscopy on water re-circulated for 50 hours through filters containing PSGO-MW and PSGO-OV powders. Detection limit, estimated with calibrated GO solutions, was 0,5-1 ppm.

Figure 5 shows the images of the recirculated water, as well as the UV-vis spectra of the filtered water (red line) and of contaminated calibration solutions having a concentration range 0,25 -10 ppm. The comparison indicates that any GO possibly released in filtered water was below 1 ppm.

Dynamic Light Scattering (DLS) measures the auto-correlation function of photons scattered by nanoscopic objects in solution, and we used it to detect the possible presence of small particles or nanosheets released in water (see details in ESI, Figs. S4-S5S13-14). Its application to non-spherical

objects is not straightforward, but it has been successfully used for the characterization of 2-dimensional materials such as graphene and BN.⁵⁸⁻⁶⁰

DLS detection limit for GO nanosheets was 5 ppm, as determined with calibration solutions of GO with known concentration (see section 13, ESI for the DLS spectrograms and measurement details). DLS measurements performed on tap water re-circulated for 100 hours through filters containing PSGOPSU-GO-MW and PSGOPSU-GO-OV powders did not show any presence of contaminants in the size range 1 nm-10 μ m for concentrations >5 ppm, confirming that no significant amounts of GO nanosheets were released by the composite within the limit of detection range. For comparison, blank PPSU-GO composites with no MW or OV stabilization treatment would instead release large amounts of debris and contaminants when suspended in water, visible by naked eye (Figure S2 in S3, ESI). This demonstrated the efficacy of the microwave processing to stabilize, at low temperature, the composite material, while keeping its adsorption performance high. Attempts to concentrate the sample to further enhance the detection of contaminants, both by ultracentrifugation and vacuum evaporation failed (Figures S16-S17, ESI).

Conclusions

The previous sections data described above demonstrate that the proposed approach is effectively a stable and cost-effective PSU-GO composite could be obtained by fixing a thin layer (ca. 10 sheets) of GO on the PPSU granules. The process is easily up-scalable, relying only on water processing and microwaves or conventional heating activation of already produced membranes, without the need of adapting already existing production technologies and plants. In particular, microwave treatments are already extensively used on large scale in industry. The stabilization effect of the treatment appears very effective, yielding no release of GO below detection limit upon extensive testing (2L/h for 100 h) as demonstrated by two different independent techniques.

The synergic effect of GO and PSU can be demonstrated by comparing the adsorption capability of the PSU-GO composite vs. the performance of bulk GO (see also table S5 and S11 in ESI). The final effective surface area measured for Rhodamine removal by PSU-GO-MW is 143 m²/g, more than 10% of the area measured in pure GO, even if only 5% of GO is present in the composite. A similar estimate can be done for OFL, where the performance is ca. 9% of bulk GO, for the 5% loading. The excellent performance obtained in removing the target contaminants from tap water can be attributed to the unique properties of 2D materials. Both GO and PSU can capture water organic contaminants; however, their meso-structure and capture mechanism is different. While PSU features a 3D network of meso- and micro-channels, GO features a stacked structure with nanometric spacing, which changes upon molecule intercalation (see XRD data in ESI). To work at best, such GO layered structure should be fully accessible from solution, i.e. arranged as thin layers, avoiding strong aggregation or bulk clusters as could happen in bulk graphite oxide powders. The structure described here, made of thin GO 2D coatings on a mechanically stable 3D structure of PSU microchannels (figures 1 and S2) is thus ideal to ensure that all GO can actively contribute to the contaminant removal.

Remarkably, the GO coating does not modify significantly the porous structure of the PPSU. The standard measurements of surface area, based on weak physisorption of N₂ molecules in gas, gives a similar SSA before and after the coating. However, the interaction of the composite with organic dye molecules in solution gives a much higher effective area of interaction, which varies with surface chemistry and target molecule structure. Quantitative analysis indicates that the adsorption process follows a BET and Langmuir model, and that the reason of improved performance is due to the few-layers structure of the GO nanosheet, which gives a higher effective area available for adsorption vs standard bulk polymer, due to intercalation of the contaminants in between the sheets. Such process is

not observed in standard surface area measurement using adsorption of nitrogen gas and is similar to what was previously observed in layered minerals like clays. Respect to other layered materials, the GO nanosheets have better processability, being soluble in water, and a versatile, tunable surface chemistry which can attract contaminants with π - π , Van der Waals or hydrogen bonding interactions.

Such material While microwaves have been used extensively to exfoliate or functionalize graphene, their use to “fix” GO on a polymer substrate has not been reported before (patent submitted).

Such method can be applied on micro and ultrafiltration membranes already commercially available, as well as scraps deriving from their preparation, without the need of adapting industrial production systems.

The PSU-GO composite exhibits thus an original adsorption mechanism, interesting for fundamental science, as well as promising potential application in commercial filters for water purification, a topic of timely importance. Further work is ongoing to better evaluate the industrial feasibility of PS-GO based filters (patent submitted).

Acknowledgements

The research leading to these results has received funding from the European Union's Horizon 2020 research and innovation programme under GrapheneCore2 785219 – Graphene Flagship and the Swedish Research Council (project Janus 2017-04456).

Conflicts of interest

The authors declare no conflicts of interest.

References

1. Y. Zhou, O. G. Apul and T. Karanfil, *Water Res.*, 2015, **79**, 57-67.
2. Y. Shen, Q. L. Fang and B. L. Chen, *Environmental Science & Technology*, 2015, **49**, 67-84.
3. G. Ersan, O. G. Apul, F. Perreault and T. Karanfil, *Water Res.*, 2017, **126**, 385-398.
4. O. G. Apul, Q. L. Wang, Y. Zhou and T. Karanfil, *Water Res.*, 2013, **47**, 1648-1654.
5. Z. Q. Niu, L. L. Liu, L. Zhang and X. D. Chen, *Small*, 2014, **10**, 3434-3441.
6. F. Perreault, A. F. de Faria and M. Elimelech, *Chem. Soc. Rev.*, 2015, **44**, 5861-5896.
7. B. Beless, H. S. Rifai and D. F. Rodrigues, *Environmental Science & Technology*, 2014, **48**, 10372-10379.
8. L. H. Jiang, Y. G. Liu, S. B. Liu, G. M. Zeng, X. J. Hu, X. Hu, Z. Guo, X. F. Tan, L. L. Wang and Z. B. Wu, *Environmental Science & Technology*, 2017, **51**, 6352-6359.
9. F. F. Liu, J. Zhao, S. G. Wang, P. Du and B. S. Xing, *Environmental Science & Technology*, 2014, **48**, 13197-13206.
10. F. Wang, Y. Wang, W. W. Zhan, S. R. Yu, W. H. Zhong, G. Sui and X. P. Yang, *Chem. Eng. J.*, 2017, **320**, 539-548.
11. L. Gan, H. Li, L. W. Chen, L. J. Xu, J. Liu, A. B. Geng, C. T. Mei and S. M. Shang, *Colloid. Polym. Sci.*, 2018, **296**, 607-615.
12. H. C. Bi, X. Xie, K. B. Yin, Y. L. Zhou, S. Wan, L. B. He, F. Xu, F. Banhart, L. T. Sun and R. S. Ruoff, *Adv. Funct. Mater.*, 2012, **22**, 4421-4425.
13. Y. Q. Chen, L. B. Chen, H. Bai and L. Li, *Journal of Materials Chemistry A*, 2014, **2**, 13744-13744.
14. J. P. Zhao, W. C. Ren and H. M. Cheng, *J. Mater. Chem.*, 2012, **22**, 20197-20202.
15. R. Rezaee, S. Nasser, A. H. Mahvi, R. Nabizadeh, S. A. Mousavi, A. Rashidi, A. Jafari and S. Nazmara, *Journal of Environmental Health Science and Engineering*, 2015, **13**.

16. H. Q. Wu, B. B. Tang and P. Y. Wu, *Journal of Membrane Science*, 2014, **451**, 94-102.
17. J. A. Prince, S. Bhuvana, V. Anbharasi, N. Ayyanar, K. V. K. Boodhoo and G. Singh, *Water Res.*, 2016, **103**, 311-318.
18. M. Miculescu, V. K. Thakur, F. Miculescu and S. I. Voicu, *Polym. Adv. Technol.*, 2016, **27**, 844-859.
19. X. Y. Zhu, K. J. Yang and B. L. Chen, *Environmental Science-Nano*, 2017, **4**, 2267-2285.
20. J. Alam, A. K. Shukla, M. Alhoshan, L. A. Dass, M. R. Muthumareeswaran, A. Khan and F. A. A. Ali, *Adv. Polym. Tech.*, 2018, **37**, 2597-2608.
21. S. Bolisetty, M. Peydayesh and R. Mezzenga, *Chem. Soc. Rev.*, 2019, **48**, 463-487.
22. C. Santhosh, V. Velmurugan, G. Jacob, S. K. Jeong, A. N. Grace and A. Bhatnagar, *Chem. Eng. J.*, 2016, **306**, 1116-1137.
23. P. Westerhoff, P. Alvarez, Q. L. Li, J. Gardea-Torresdey and J. Zimmerman, *Environmental Science-Nano*, 2016, **3**, 1241-1253.
24. M. P. Wei, H. Chai, Y. L. Cao and D. Z. Jia, *J. Colloid Interface Sci.*, 2018, **524**, 297-305.
25. N. Oger, Y. T. F. Lin, C. Labrugere, E. Le Grogne, F. Rataboul and F. X. Felpin, *Carbon*, 2016, **96**, 342-350.
26. J. Abraham, K. S. Vasu, C. D. Williams, K. Gopinadhan, Y. Su, C. T. Cherian, J. Dix, E. Prestat, S. J. Haigh, I. V. Grigorieva, P. Carbone, A. K. Geim and R. R. Nair, *Nature Nanotechnology*, 2017, **12**, 546-+.
27. C. J. Madadrang, H. Y. Kim, G. H. Gao, N. Wang, J. Zhu, H. Feng, M. Gorrington, M. L. Kasner and S. F. Hou, *Acs Applied Materials & Interfaces*, 2012, **4**, 1186-1193.
28. W. Gao, M. Majumder, L. B. Alemany, T. N. Narayanan, M. A. Ibarra, B. K. Pradhan and P. M. Ajayan, *Acs Applied Materials & Interfaces*, 2011, **3**, 1821-1826.
29. Y. Shen and B. L. Chen, *Environmental Science & Technology*, 2015, **49**, 7364-7372.
30. W. K. Chee, H. N. Lim, N. M. Huang and I. Harrison, *Rsc Advances*, 2015, **5**, 68014-68051.
31. M. Zambianchi-nchi, M. Durso, A. Liscio, E. Treossi, C. Bettini, M. L. Capobianco, A. Aluigi, A. Kovtun, G. Ruani, F. Corticelli, M. Brucale, V. Palermo, M. L. Navacchia and M. Melucci, *Chem. Eng. J.*, 2017, **326**, 130-140.
32. M. Zambianchi, A. Aluigi, M. L. Capobianco, F. Corticelli, I. Elmi, S. Zampolli, F. Stante, L. Bocchi, F. Belosi, M. L. Navacchia and M. Melucci, *Advanced Sustainable Systems*, 2017, **1**, #1700019.
33. T. Laaninen, *European Parliamentary Research Service*, 2018, **PE 625.179**.
34. V. Geissen, H. Mol, E. Klumpp, G. Umlauf, M. Nadal, M. van der Ploeg, S. van de Zee and C. J. Ritsema, *International Soil and Water Conservation Research*, 2015, **3**, 57-65.
35. S. Schulze, D. Zahn, R. Montes, R. Rodil, J. B. Quintana, T. P. Knepper, T. Reemtsma and U. Berger, *Water Res.*, 2019, **153**, 80-90.
36. D. Pierleoni, Z. Y. Xia, M. Christian, S. Ligi, M. Minelli, V. Morandi, F. Doghieri and V. Palermo, *Carbon*, 2016, **96**, 503-512.
37. D. Pierleoni, M. Minelli, S. Ligi, M. Christian, S. Funke, N. Reineking, V. Morandi, F. Doghieri and V. Palermo, *ACS Applied Materials & Interfaces*, 2018, **10**, 11242-11250.
38. V. Palermo, I. A. Kinloch, S. Ligi and N. M. Pugno, *Adv. Mater.*, 2016, **28**, 6232-6238.
39. M. Melucci, E. Treossi, L. Ortolani, G. Giambastiani, V. Morandi, P. Klar, C. Casiraghi, P. Samori and V. Palermo, *J. Mater. Chem.*, 2010, **20**, 9052-9060.
40. A. Kovtun, D. Jones, S. Dell'Elce, E. Treossi, A. Liscio and V. Palermo, *Carbon*, 2019, **143**, 268-275.
41. A. J. Ebele, M. Abou-Elwafa Abdallah and S. Harrad, *Emerging Contaminants*, 2017, **3**, 1-16.
42. K. X. Sun, S. N. Dong, Y. Y. Sun, B. Gao, W. C. Du, H. X. Xu and J. C. Wu, *J. Hazard. Mater.*, 2018, **348**, 92-99.
43. H. Y. Guo, T. F. Jiao, Q. R. Zhang, W. F. Guo, Q. M. Peng and X. H. Yan, *Nanoscale Research Letters*, 2015, **10**.

44. K. S. W. Sing, D. H. Everett, R. A. W. Haul, L. Moscou, R. A. Pierotti, J. Rouquerol and T. Siemieniewska, *Pure Appl. Chem.*, 1985, **57**, 603-619.
45. A. Ebadi, J. S. S. Mohammadzadeh and A. Khudiev, *Adsorption-Journal of the International Adsorption Society*, 2009, **15**, 65-73.
46. A. Schlierf, H. F. Yang, E. Gebremedhn, E. Treossi, L. Ortolani, L. P. Chen, A. Minoia, V. Morandi, P. Samori, C. Casiraghi, D. Beljonne and V. Palermo, *Nanoscale*, 2013, **5**, 4205-4216.
47. M. J. McAllister, J. L. Li, D. H. Adamson, H. C. Schniepp, A. A. Abdala, J. Liu, M. Herrera-Alonso, D. L. Milius, R. Car, R. K. Prud'homme and I. A. Aksay, *Chem. Mater.*, 2007, **19**, 4396-4404.
48. J. C. Santamarina, K. A. Klein, Y. H. Wang and E. Prencke, *Canadian Geotechnical Journal*, 2002, **39**, 233-241.
49. G.-J. Su, S.-X. Yin, L.-J. Wan, J.-C. Zhao and C.-L. Bai, *Surf. Sci.*, 2004, **551**, 204-212.
50. J. T. J. Freitas, C. C. de Melo, O. Viana, F. F. Ferreira and A. C. Doriguetto, *Crystal Growth & Design*, 2018, **18**, 3558-3568.
51. V. Palermo and P. Samori, *Angew. Chem. Int. Ed.*, 2007, **46**, 4428-4432.
52. E. Treossi, M. Melucci, A. Liscio, M. Gazzano, P. Samorì and V. Palermo, *J. Am. Chem. Soc.*, 2009, **131**, 15576-15577.
53. R. R. Nair, H. A. Wu, P. N. Jayaram, I. V. Grigorieva and A. K. Geim, *Science*, 2012, **335**, 442-444.
54. G. Romanelli, A. Liscio, R. Senesi, R. Zamboni, E. Treossi, F. Liscio, G. Giambastiani, V. Palermo, F. Fernandez-Alonso and C. Andreani, *Carbon*, 2016, **108**, 199-203.
55. Medica, <https://www.medica.it/>.
56. A. Bianco, *Angew. Chem. Int. Ed.*, 2013, **52**, 4986-4997.
57. J. Zhao, Z. Y. Wang, J. C. White and B. S. Xing, *Environmental Science & Technology*, 2014, **48**, 9995-10009.
58. K. Kouroupis-Agalou, A. Liscio, E. Treossi, L. Ortolani, V. Morandi, N. M. Pugno and V. Palermo, *Nanoscale*, 2014, **6**, 5926-5933.
59. A. Catheline, L. Ortolani, V. Morandi, M. Melle-Franco, C. Drummond, C. Zakri and A. Penicaud, *Soft Matter*, 2012, **8**, 7882-7887.
60. M. Lotya, A. Rakovich, J. F. Donegan and J. N. Coleman, *Nanotechnology*, 2013, **24**, #265703.