

# **ClimAg model description** Version 2.0

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# **1. Introduction**

The ClimAg model is a biophysical systems model that calculates the resource use and emissions of greenhouse gases (GHGs) and nitrogen pollutants from food, fiber, and biofuel production. The primary application of ClimAg is to calculate the climate impact of food and biofuel production systems. In addition to recurring GHGs, the model calculates the climate impact of carbon stock changes in plants and soils caused by land use.

ClimAg models all major steps related to agriculture and aquaculture production and use of food, materials, and biofuels, including: i) production of inputs (fertilizer, electricity, etc.); ii) crop, livestock, and seafood production; iii) processing into end-use-ready items; iv) end-use (consumption); and v) transportation between production and use nodes. The model also represents all major co-products and their use; see Figure 1.



**Figure 1** Overview of the ClimAg model system: Sub-systems included and major product and co-product flows. Some flows are indicated for clarity. Emission flows are not shown. Sub-systems not shown are freight transport and production of fuels, electricity, fertilizer, and pesticides.

Key design features of the ClimAg model include:

 **Consistent accounting of upstream resource use and emissions** of all feeds and feedstocks used in production systems. The ClimAg model consistently calculates the land and energy use, and GHG and nitrogen (N) emissions that occur in the supply of all categories of feeds and feedstocks (see Table 23). This applies also to all flows generated as co-products, e.g., cereal brans and oil meals. Such upstream costs are also calculated for co-products, which are often considered free in

other models and analyses. For example, straw used for bioenergy or manure used for organic crop production are typically assigned no upstream climate cost.

- **Physically consistent representation of the production and use of co-products** generated in crop and livestock systems, and related processing industries (see Table 16-22 for co-products included). Most co-products are useful as feedstock in other production processes. ClimAg calculates the production of co-products based on mass- and energy-balanced descriptions of the processes in which they originate. This ensures that the availability of co-products is correctly scaled to the production levels in the sub-systems that generate the co-products.
- **Endogenous representation of livestock herds** in terms of number of animals of different functions and ages, and the herd output of milk/egg and slaughter animals. Herd size and structure are calculated using herd dynamics parameters (e.g., reproduction and growth rates, and animal cohort descriptions, mainly age and liveweight); see section 3.1.1. Endogenous representation enables calibration of key herd productivity parameters, such as calving and liveweight gain rates, to country statistics on production per number of livestock.
- **Endogenous estimates of feed energy intake per animal**, calculated with empirically based equations that use various herd characteristics parameters as input data (e.g., liveweight, growth rate, and milk/egg production rate);see 3.2.1. Endogenous calculations of feed energy intake ensure fairly accurate feed use estimates even when feed basket data are incomplete. The benefit of this model feature is particularly important in systems with significant grazing, as the amount of grazed feed is rarely known.
- **Description of nitrogen (N) flows on a mass balance basis**. The ClimAg model includes a highly detailed, mass-balance based representation of N flows in the food and agriculture system (see Figure 3). Mass-balanced descriptions of N flows improve the accuracy of emission estimates for crop and livestock production, from which substantial amounts of N can escape as different gases and nitrate. Most of these losses are expensive to measure directly and rarely known with high certainty. Using mass balance ensures physically consistent results, and more accurate estimates overall of N flows.

The model calculates land and energy use, climate impacts, and N emissions for approximately 400 products and by-products from agriculture, aquaculture, and fisheries (see Table 16-22) and includes most GHG emission sources from agriculture and aquaculture:

- $\bullet$  Nitrous oxide (N<sub>2</sub>O) from mineral soils, with separate representation of emissions from:
	- Plant residues left in field, including root mass
	- Fertilizer application, specific to crop type
	- Manure application, specific to crop type, manure type, and application technology
	- Manure excreted at grazing
- $\bullet$  CO<sub>2</sub> and N<sub>2</sub>O from drained organic soils
- $\bullet$  Methane (CH<sub>4</sub>) from flooded rice fields
- CH<sub>4</sub> from feed digestion ("enteric" fermentation) in ruminants and pigs
- $\bullet$  N<sub>2</sub>O and CH<sub>4</sub> from livestock manure in animal confinements and storage, respectively
- CH<sub>4</sub> from manure excreted at grazing
- CH<sub>4</sub> and N<sub>2</sub>O from aquaculture facilities
- "Indirect"  $N_2O$  caused by ammonia and nitrate emissions from agriculture
- $\bullet$  CO<sub>2</sub> from fuel and electricity use in crop production (e.g., for land preparation, irrigation, harvesting, and post-harvest crop drying).
- CO<sub>2</sub> from fuel and electricity use in livestock confinements, aquaculture facilities, and capture fisheries
- $\bullet$  CO<sub>2</sub> from fuel and electricity use in crop, livestock, and fish/shellfish processing
- $\bullet$  CO<sub>2</sub> and N<sub>2</sub>O from production of mineral fertilizers and pesticides
- $\bullet$  CO<sub>2</sub> from manufacturing of materials used in greenhouse structures
- CO<sub>2</sub> from transportation from production to use and end-use, including inter-regional trade

ClimAg does not include energy use and emissions from manufacturing and construction of machinery, buildings, etc., except for greenhouse structures. In addition, GHG emissions from packaging, retailing or food preparation are not included. Lastly, ClimAg does not include CO<sub>2</sub> emissions from liming, partly because it is a very small source, and partly because the net GHG emissions due to liming are very uncertain and may even be negative (Karlsson Potter et al., 2023).

In addition to the recurring GHG emission sources listed above, ClimAg includes the climate impact of carbon stock changes in plants and soils caused by land use, here referred to as the "carbon opportunity cost" of land use (see 2.5).

When vegetation changes from forest to agricultural land, and vice versa, changes occur in surface albedo and cloud formation which both influence regional and global temperatures. However, due to insufficient understanding and data in the literature, ClimAg does not include these factors.

As mentioned above, production systems included in ClimAg are described on a nitrogen mass-balance basis. The model includes all system inputs of fixed N (fertilizer, biological fixation, and atmospheric deposition), and all potential N losses (emissions). These emissions include:

- Ammonia ( $NH<sub>3</sub>$ ) from soils, with separate representation of emissions from:
	- Fertilizer application on cropland, specific by manure type
	- Manure application on cropland, specific by manure type
	- **Manure excreted at grazing, specific by urine and feces, and by species**
	- Decomposing crop residues left in field after harvest
- Nitric oxide (NO) from soils
- Dinitrogen  $(N_2)$  from soils
- Nitrate  $(NO<sub>3</sub>^-)$  from soils
- NH<sub>3</sub> from livestock confinements and manure storage facilities
- Runoff of N from livestock confinements and manure storage facilities

# **2. Crop and pasture production**

Agricultural land use for food, fiber, and biofuel production is very diverse. To capture the variation that is relevant to resource and environmental impacts, many types of crops and land use systems have been included in the ClimAg model. In total about 70 different annual and perennial crops on cropland are represented, see Table 16. Between them, these crops account for close to 100% of global cropland use. In addition to crops cultivated in the open field, ClimAg also includes four crops grown in greenhouses: tomatoes, cucumbers, peppers (capsicum), and eggplants.

Pasture production on permanent and semi-permanent grassland is represented by four types, based on the pre-existing native vegetation on the site: originally forest, originally tropical/sub-tropical grassland or woodland, originally temperate grassland, and originally xeric grassland. This distinction helps to facilitate estimates of the climate and environmental impact of permanent and semi-permanent pastures. Semi-permanent grassland refers to grassland that is renewed every 10-15 years by plowing and sowing. This type of grassland is relatively common in, for example, Brazil.

Where applicable, separate representation is done for conventional and organic systems, respectively. The model does not include separate representation of rainfed and irrigated production, except for the calculations of carbon opportunity cost, see 2.5.

Table 1-3 list main exogenous parameters in crop and pasture systems. The sections below provide more details.

# **2.1 Plant growth and land areas**

### *2.1.1 Plant growth and crop production*

Using the parameters in Table 1, the model calculates the total net photosynthetic production per hectare and year and its partition into different plant components (harvestable items, roots etc).

### *Annual crops*

For annual crops on arable land, total net photosynthetic production per hectare is calculated as a function of net harvested mass ('yield'), harvest losses, and allometric relationships of plant organs (harvestable part, non-harvested parts above ground, root mass). For most cereals, oil/protein crops, and starchy root crops, the fraction harvestable part ('harvest index'),  $\alpha^c$ , is calculated endogenously as a function of yield, using the equations developed by (García‐Condado et al., 2019).

Net plant mass production per hectare and year for crop  $c$ ,  $P_{net}^c$  (Mg dry matter ha<sup>-1</sup> yr<sup>-1</sup>), is thus calculated according to:

$$
P_{net}^c = \frac{gross \, yield \times dry \, matter \, content \, of \, yield}{\alpha^c} \times \frac{1}{(1-\% \, root \, mass \, of \, total \, plant \, mass)} \quad \text{Eq. 1}
$$

where:

*gross yield* = *net yield* + *harvest spill/losses* (see Table 1)

Numbers on dry matter content of yield and percentage root mass are given in Table 25.

**Table 1** Main exogenous parameters in crop and pasture sub-systems. Parameters describing nitrogen flows are shown in Table 3. Parameters describing plant growth for tree and bush crops are shown in Table 2. Sub-systems are listed in Table 16.





#### *Perennial grasses and legumes*

For perennial grasses and/or legumes cultivated on arable land, and permanent and semi-permanent grasslands, growth above ground is stated exogenously. This is done separately for two seasons: the wet and/or warm season, which is the main grazing season, and the dry and/or cold season. Root mass production (including exudates) is calculated as an allometric relationship to above-ground plant mass and a root mass turnover rate.

#### *Perennial trees and bushes*

Perennial tree and bush crops are represented in ClimAg by 17 different sub-systems (see list in Table 16). Plantains and bananas are included in this group, although they are not trees nor bushes, and almost all their above-ground plant mass is annual (only the corm and roots are perennial).

The representation of the production cycle includes a separate establishing phase, during which the planted saplings grow but do not yield harvestable products. The establishing phase is typically 2-6 years depending on species. The production phase varies from about 10 to 60 years depending on species.

Plant growth is calculated as the annual turnover of perennial plant mass, and as the annual growth of the harvested part and other annual plant mass, see Table 2 for parameters. Perennial plant mass includes roots, corm, trunk, twigs & branches, and leaves on evergreens (oil palm, coconut, olive, cashew, mango, orange, cocoa, coffee, tea). Annual plant mass other than the product includes leaves on deciduous trees and bushes (almond, grapes, apple) and above-ground growth of banana and plantain. Plant mass stocks of the entire plant are stated exogenously, at the start of the establishing phase and at the end of the production phase, respectively. Exogenous allometric parameters are used to calculate stocks of individual plant components (roots, trunk, twigs/branches/corm, and leaves).

### *2.1.2 Land areas*

In the ClimAg model, a distinction is made between "physical" land area and "harvested" area. The physical area is the land area occupied by a crop, pasture, or other land use (e.g., fallow). On the physical area, one or more harvests can take place during the cultivation cycle (e.g., by double-cropping), with the sum of the harvests being the harvested area.

For permanent/semi-permanent pastures and tree/bush crops, the model calculates the required physical area simply as a function of grazed intake or net harvested yield and the required production on the farm, region, or country, etc.

For arable land, the calculation of the required area is more complex. For arable crops, the physical area of each crop is calculated in a crop-rotation framework, with the number and length of cultivation cycles per rotation cycle as exogenous variables. This allows for the modeling of so-called double cropping rotations, in which a sequence of two crops is grown in the same growing season, for example, rice-rice or soybeanmaize. For grass-legume crops on arable land, multiple harvests per year can be modeled.



**Table 2** Main exogenous parameters that describe plant growth for tree and bush crops.. Sub-systems are listed in Table 16. Parameter default values are given in Table 24.

### **2.2 Nitrogen inputs and emissions**

In crop and pasture systems, nitrogen (N) flows are described with explicit distinction between inorganic N and organic N. All processes are described on a N mass balance basis.

For inorganic N in the soil-plant system, a further distinction is made between N that potentially can be taken up by the growing plant, here referred to as "up-takeable" N, and N that cannot. For example, the addition of fertilizer-nitrogen during the growth period is considered potentially up-takeable, whereas the atmospheric deposition of N outside the growing season is not.

The representation of the N flows and stocks in the soil-plant system assumes steady-state conditions, i.e., soil N pools are assumed to be constant. This means that the annual amount of soil organic N converted to mineral N equals the annual input to the soil of plant matter and organic manure N.

**Table 3** Main exogenous parameters for nitrogen (N) flows in crop and pasture systems.





#### *2.2.1 Nitrogen balance of the soil-plant system*

Using the description of total net photosynthetic production (see 2.1.1) as a basis, the model first calculates the *total net nitrogen assimilation* in crop and pasture production,  $N_{ass}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>). This is calculated as the growth of different plant components multiplied by their N content (protein content divided by 6.25, see Table 25 for numbers).

Next, the model calculates the *required supply of up-takeable nitrogen* by the growing plant,  $N_{\text{req updateable}}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>). The required supply for up-takeable N of crop *c* is calculated as the total net N assimilation of the crop ( $N_{ass}^c$ ) divided by the efficiency of uptake of N available for uptake,  $\varepsilon_{ass}^c$ :

$$
N_{req\ uptakeable}^{c} = \frac{N_{ass}^{c}}{\varepsilon_{ass}^{c}}
$$
 Eq. 2

The parameter  $\varepsilon_{ass}^c$  (dimensionless) reflects differences between crop species in the uptake efficiency of soil N related to their varying density and depth of root systems (see Velthof et al. 2009). For example, perennial grasses have very extensive root systems and, therefore, have high uptake efficiencies (close to 100%). However, most other crops have less extensive systems and are therefore unable to take up close to 100% of the plant-available N in the root zone.

Next, the model calculates all inputs. These inputs are i) decomposing organic N in plant matter left in the field from preceding crops, ii) biological fixation, iii) manure, iv) mineral fertilizer, and v) atmospheric deposition. After this, ammonia emissions associated with these inputs are calculated. The remaining nitrogen after ammonia losses is here defined as the *actual net supply* of N,  $N_{net\,sup}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), to the soil:

$$
N_{net\,sup}^c = N_{res}^r + N_{fix}^c + N_{man}^c + N_{syn}^c + N_{dep}^c - N_{NH3}^c
$$
 Eq. 3

where:

 $N_{res}^{r}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) the total N supply in decomposed plant mass from the previous crop (see 2.2.2)  $N_{fix}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the total N input through biological fixation (see 2.2.3)  $N_{man}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the total N in manure excreted or applied on field (see 2.2.4)  $N_{syn}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the total N input through mineral fertilizer (see 2.2.5)  $N_{den}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the total N input through atmospheric deposition  $N_{NH_2}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) are the total emissions of ammonia (see 2.2.6)

Then the model calculates how much of this net supply can contribute to the *supply of up-takeable* nitrogen  $N_{sup}^c$  uptakeable (kg N ha<sup>-1</sup> yr<sup>-1</sup>), by applying a parameter, here denoted  $\eta_f^c$  (dimensionless), that reflects inefficiencies in N utilization related to mismatches in supply and uptake:

$$
N_{sup\ uptakeable}^c = \sum_{f=1}^n \eta_f^c \times N_{net\ sup f}^c
$$
 Eq. 4

where:

 $\eta_f^c$  (dimensionless) is the fraction of net N supply from source f (see Eq. 3) taken up by crop c

 $N_{net\,sun\,f}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) the net N supply from source f

For several reasons, in real systems, the net supply is almost always larger than the potential uptake. This is because, for example, the mineralization of organic N and atmospheric deposition both occur during the non-growing season. This parameter is also applicable to the supply of manure and fertilizer, which typically are supplied in excess of the potential uptake. This is partly because of local excess supply, in the form of urine patches and manure being applied outside the growing period. More importantly, it is due to fundamental uncertainties in estimating uptake, because of the impossibility of predicting weather conditions over the entire growing period. Faced with this uncertainty, farmers tend to anticipate higher yields and, accordingly, apply fertilizer and manure at the higher end.

Lastly, the model calculates the net plant-soil nitrogen balance, as the sum of all external N inputs minus N removed in harvested or grazed plant mass,  $N_{rem}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), and all gas outputs except dinitrogen  $(NH<sub>3</sub>, N<sub>2</sub>O, NO)$ . In most systems, the balance is positive, i.e., the inputs exceed the outputs. This surplus is assumed to leave the soil as nitrate and dinitrogen (see 2.2.7).

#### $2.2.2$ Nitrogen supply from previous crop

The N net supply from decomposing crop residues and other plant matter left in the field from preceding crops is calculated as the N content of plant mass left in the field after harvest or grazing period, minus emissions of ammonia that occur during the decomposition process (less than 5% of total above-ground organic N is typically lost).

It should be noted that the decomposition of plant matter left in the field takes several years to be complete. Here, steady-state conditions are assumed, where the annual amount of plant matter left in the

field is the same over several years. In this way, the annual amount of organic N converted to mineral N equals the annual input from decomposing plant matter.

For arable-land crops, the net N supply in decomposed plant mass from the previous crop,  $N_{res}^r$  (kg N ha<sup>-1</sup>  $yr^{-1}$ ), is calculated as an area-weighted average for all crops,  $c = 1, 2, ..., n$ , in the crop rotation, r.

$$
N_{net\ res}^r = \sum_{c=1}^n \left( N_{lft}^c \times \beta_r^c - N_{lft,abo}^c \times \%NH_{3lf,abo}^c \times \beta_r^c \right)
$$
 Eq. 5

where:

 $\beta_r^c$  (dimensionless) is the average arable-land area fraction of crop c over the *entire crop rotation* cycle.

 $N_{tft}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the total N content of plant mass for crop c left in field after harvest or grazing period.

 $N_{lft,abo}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the *above-ground* N content of plant mass for crop c left in field after harvest and/or uneaten plant mass at grazed areas.

%NH3<sup>c</sup><sub>Ift,abo</sub> (dimensionless) is the fraction of above-ground N in plant mass for crop c lost through ammonia volatilization.

For example, for a 4-yr crop rotation consisting of a 3-yr grass ley and a 1-yr barley crop,  $\alpha_r^{ley}$  would be 75% and  $\alpha_r^{barley}$  25%. Assuming that  $N_{lft}^{ley}$ =140 and  $N_{lft}^{barley}$  = 40, and %NH3 $c_{lft,abo}^c$  = 0 for simplicity, then  $N_{res}^r = 115$  kg N ha<sup>-1</sup> yr<sup>-1</sup>. This number, hence, represents the *annual average* pre-crop N supply *at* steady state for all crops in the rotation.

For permanent grasslands, and tree and bush crops, net N supply is simply calculated as the N content in decomposing plant mass left in field from the previous growth cycle, minus ammonia losses.

The up-takeable fraction,  $\eta_{res}^c$ , of nitrogen from decomposed plant matter depends on how much of the decomposition occurs under periods with crop uptake of N. As a default, ClimAg uses 70% for annual crops, and 90% for permanent grassland, based on Velthof et al. (2009).

#### $2.2.3$ Nitrogen net supply from biological fixation

For crops in symbiosis with nitrogen-fixing bacteria, N addition to the soil by fixation,  $N_{fix}^c$  (kg N ha<sup>-1</sup> yr <sup>1</sup>), is calculated as a percentage,  $\%N_{dfa}$ , of the total N in the annual net photosynthetic production,  $N_{ass}^c$ , plus N transferred to co-growing crops and/or immobilized. This is in line with other models (see, for example, Høgh-Jensen et al., 2004):

$$
N_{fix}^c = \frac{96}{M_{dfa}} \times N_{ass}^c \times (1 + \frac{96}{M_{trans}})
$$
 Eq. 6

where  $\%N_{trans}$  (dimensionless) is the fraction of N transferred to co-growing crops and/or immobilized.

For crops with plant-associated but non-symbiotic nitrogen-fixing bacteria,  $N_{fix}^c$  is simply an exogenously stated quantity.

For symbiotic fixation, no losses are assumed in the uptake by the plant, since the fixed N is fed directly into the plant roots. However, the amount of fixed N deposited in the soil (biologically fixed N transferred and/or immobilized) enters the soil N pool and is, therefore, subject to losses before being taken up by the subsequent crops. Furthermore, N in plant matter from nitrogen-fixing crops left in field is subject to losses described in 2.2.2.

#### $2.2.4$ Nitrogen supply from manure

The calculation of the N supply from manure represents the organic and inorganic N fractions in the manure separately. It also calculates the supply separately from the land application of manure excreted in confinements and manure excreted during grazing. For both applied and excreted manure, the N net supply is calculated as the N content of the manure, minus emissions of ammonia that occur after application or excretion.

The amounts and N contents of manure from confinements and manure excreted during grazing are calculated endogenously in ClimAg, see 3.3.1 and 3.3.4. For manure from confinements applied on cropland, the net supply of inorganic and organic N  $N_{net~st.man,ino}^c$  and  $N_{net~st.man,org}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), is calculated as the amount of manure divided by the area of cropland (ha) receiving manure, minus ammonia losses:

$$
N_{net\;st.man,ino}^{c} = \frac{\% \, applied \times N_{man,sto,ino}^{l}}{cropland\;area \, applied} \times (1 - \%NH_{3st.man,ino}^{c,t,a,s})
$$
 Eq. 7

$$
N_{net\;st.man,org}^{c} = \frac{\% \, applied \times N_{man,sto,ino}^{l}}{cropland\;area\;applied} \times (1 - \%NH_{3st.man,org})
$$
 Eq. 8

where:

% applied is the fraction of manure produced at the livestock farm applied on cropland at the farm. If not applied, the manure is sold, burnt, or otherwise lost.

 $N_{man, sto, ino}^{l}$  (kg N yr<sup>-1</sup>) is output from manure storage at the livestock farm of inorganic manure N, see Eq. 56.

 $N_{man,sto,org}^{l}$  (kg N yr<sup>-1</sup>) is output from manure storage at the livestock farm of organic manure N, see Eq. 55.

% $NH_3^{c,t,a,s}$  (dimensionless) is the fraction of N input for crop c lost through ammonia volatilization. The index  $t$  denotes the sub-type of input (different types of manure), the index  $a$  denotes the sub-set of application technologies, and the index  $s$  the time (season) of manure application (see Table 3).

The calculation of the N supply from manure excreted during grazing is done in a way similar to that of stored manure. All excreted manure is assumed to be left on the pasture, except in cases of manure collection for use as fuel, etc.

The amount of up-takeable N from inorganic and organic manure N is calculated separately for i) applied manure on cropland, ii) excreted feces on pasture, and iii) excreted urine on pasture, with explicit distinction between inorganic and organic N for all of them.

For organic manure N applied on cropland, default values on the up-takeable fraction,  $\eta_{st, man, ora}^c$ , are the same as those for decomposed plant matter (see 2.2.2). For inorganic N, the up-takeable fraction of inorganic N,  $\eta_{st.man,ino}^c$ , depends on the timing of application and the degree of application in excess of uptake. Application outside the crop growth period is associated with a very low up-takeable fraction.

For excreted feces and urine, the fraction is somewhat lower because of the uneven spatial distribution with a high input at urine and feces patches (Velthof et al. (2009).

#### $2.2.5$ Nitrogen supply from fertilizer

Formally, the input of N fertilizer is calculated as the balance between the required supply by the crop and the net inputs from all other sources. That is, any deficit of up-takeable soil nitrogen in relation to the amount required by the crop is met by supply of fertilizer nitrogen,  $N_{syn}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>):

$$
N_{syn}^{c} = \frac{1}{(1 - \gamma_0 N H 3_{syn}^{c, t, a}) \times \eta_{syn}^{c}} \times [N_{req\ uptakeable}^{c} - N_{net\ res}^{r} \times \eta_{res}^{c} - N_{fix}^{c} - N_{dep}^{c} \times \eta_{dep}^{c} - N_{dep}^{c} \times \eta_{dep}^{c} - N_{net\ str, man, ino}^{c} \times \eta_{str, man, ino}^{c} - N_{net\ gr. vir, ino}^{c} \times \eta_{gr. vir, ino}^{c} - N_{net\ gr. rec, ino}^{c} \times \eta_{gr. rec, ino}^{c} - N_{net\ str, man, org}^{c} \times \eta_{str, man, org}^{c} - (N_{net\ gr.uri, org}^{c} + N_{net\ gr. fec, org}^{c}) \times \eta_{gr. man, org}^{c} ]
$$
 Eq. 9

where:

 $N_{\text{req uptakeable}}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the required supply of up-takeable N by crop c, see Eq. 2.

 $N_{net}^{r}$  res (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the net (i.e., after ammonia losses) N supply from decomposed plant matter left in field from the previous crop, see 2.2.2.

 $N_{fix}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the input through biological N fixation, see 2.2.3.

 $N_{dep}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the input through atmospheric deposition.

 $N_{net\;st.max,org}^c$  and  $N_{net\;st.max,ino}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the net (i.e., after ammonia losses) supply of organic and inorganic N, respectively, in manure from confinements.

 $N_{net, gr. fec, org}^{c}$ ,  $N_{net, gr.uri, org}^{c}$ ,  $N_{net, gr. fec, ino}^{c}$  and  $N_{net, gr.uri, ino}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the net (i.e., after ammonia losses) supply of organic and inorganic N in feces and urine, respectively, excreted during grazing.

 $\eta_f^c$  (dimensionless) is the fraction of net input f that contributes to the supply of up-takeable N for crop  $c$ , see 2.2.1.

The fraction of up-takeable N from applied fertilizer nitrogen,  $\eta_{syn}^c$ , can, in theory, be close to 100%. In practice, however, application in excess of uptake is prevalent, leading to a lower up-takeable fraction and lower efficiency in fertilizer use.

#### Emissions of ammonia, nitrogen dioxide, and nitric oxide 2.2.6

Losses of ammonia,  $N_{NH_3}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), for crop c, are calculated according to:

$$
N_{NH_3}^c = N_{syn}^c \times \%NH_{3syn}^{c,t,a} + N_{st.man,org}^c \times \%NH_{3st.man,org}^{c,s} + N_{st.man,ino}^c \times \%NH_{3st.man,ino}^{c,t,a,s}
$$
  
\n
$$
(N_{gr. fec,org}^c + N_{gr.uri,org}^c) \times \%NH_{3gr.man,org}^c + N_{gr. fec,ino}^c \times \%NH_{3gr. fec.man,ino}^{c,t,a,s} + N_{gr.ire.man,ino}^c \times \%NH_{3gr. fec.man,ino}^{c,t,a,s}
$$
  
\nEq. 10  
\nEq. 10

where:

 $N_{syn}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the input through mineral fertilizer.

 $N_{st, man, org}^c$  and  $N_{st, man, ino}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) are the organic and inorganic N, respectively, in stored manure (from pens/confinements) applied to crop *c*.

 $N_{gr. fec, org}^c$ ,  $N_{gr.uri, org}^c$ ,  $N_{gr. fec, ino}^c$  and  $N_{gr.uri, ino}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) are the organic and inorganic N in feces and urine, respectively, excreted during grazing.

 $N_{tft,abo}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the above-ground N content of plant mass left in field after harvest and/or uneaten plant mass at grazed areas.

% $NH_{3f}^{c,t,a,s}$  (dimensionless) is the fraction of N input *f* for crop *c* lost through ammonia volatilization. The index *t* denotes the sub-type of input (different types of fertilizer or stored manure), the index *a* denotes the sub-set of application technologies, and the index *s* denotes the time (season) of manure application (see Table 3).

For the emission factor for fertilizer,  $\%NH_{3syn}^{c,t,a}$ , there are separate entries for each of the five types of mineral fertilizer *t* included in the model. In addition, there are separate values for the case where the fertilizer is incorporated in the soil and the case where it is not. Hence, there are ten different  $\%NH_{3sym}$  in the model.

Similarly, for the emission factors for *inorganic* N in *stored* manure (from pens/confinements), i.e. %NH<sub>3st.man,ino</sub>, there are separate entries depending on manure type *t* (see Table 6), application technology *a* (see Table 1) and time of application *s*. For each manure type, there are 12 different % $NH_{3st. man, ino}^{c}$  in the model.

Emissions of *nitrous oxide*,  $N_{N20}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), for crop *c*, are calculated according to:

$$
N_{N_2O}^c = N_{net\ syn}^c \times \%N_2O_{syn}^{c,t} + N_{net\ str,man}^c \times \%N2O_{st,man}^{c,t,a} + N_{net\ gr,man}^c \times \%N_2O_{gr,man}^{c,t} + N_{lft,abo}^c \times (1 - \%NH_{1ft,abo}^c) \times \%N_2O_{lt,abe}^c + N_{lf,bel}^c \times \%N_2O_{lt,bel}^c
$$
 Eq. 11

where:

 $N_{net, syn}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the net (i.e., after ammonia losses) supply through mineral fertilizer.

 $N_{net st. man}^c$  kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the net (i.e., after ammonia losses) supply of N in manure from confinements.

 $N_{net\,gr.man}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the net (i.e., after ammonia losses) supply of N in feces and urine excreted during grazing.

 $N_{tft,abo}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the above-ground N content of plant mass left in field after harvest and/or uneaten plant mass at grazed areas.

 $N_{I}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the N content of below-ground (root) plant mass remaining after harvest and/or annual turnover of root mass in perennial crops (grasses or tree and bush crops).

% $NH_3^c$  (dimensionless) is the fraction of N input *f* for crop *c* lost through ammonia volatilization.

% $N_2O_f^{c,t,a}$  (dimensionless) is the fraction of N input *f* for crop *c* emitted as nitrous oxide. The index *t* denotes the sub-type of input (different types of fertilizer or stored manure, or different types of grazing manure), and the index *a* denotes the sub-set of *manure* application technologies

It should be noted that, as described in the equation above, the emission factors  $\%N_2O_f^{c,t,a}$  are applied to the net supply of each input, i.e., after deducting ammonia losses.

In addition to variation by crop, there are separate entries for  $\frac{6}{N_2O_f^{c,t,a}}$  for different types of stored manure (liquid or solid) and different types of manure application technologies (surface application or injection/incorporation).

The calculation of *nitric oxide* emissions,  $N_{NO}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), is done in a manner analogous to that of nitrous oxide, but with no differentiation between fertilizer/manure types or application technologies.

#### $2.2.7$ Emissions of nitrate and dinitrogen

As mentioned in 2.2.1, plant-soil nitrogen balance,  $N_{\text{solid+pl,bal}}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>), for crop c is calculated as the sum of all external N inputs minus N harvested or grazed and all gas outputs except dinitrogen:

$$
N_{solid+pl,bal}^{c} = N_{syn}^{c} + N_{st.man}^{c} + N_{gr.man}^{c} + N_{fix}^{c} + N_{dep}^{c} - N_{rem}^{c} - N_{NH_3}^{c} - N_{N_2O}^{c} - N_{NO}^{c}
$$
 Eq. 12

where:

 $N_{syn}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the input through mineral fertilizer.

 $N_{st,man}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the nitrogen in stored manure (from pens/confinements) applied.

 $N_{\text{arman}}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the nitrogen in feces and urine excreted during grazing.

 $N_{fix}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the input through biological fixation.

 $N_{dep}^{c}$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the input through atmospheric deposition.

 $N_{rem}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) is the nitrogen in plant mass removed through harvest and/or grazing.

 $N_{NH_3}^c$ ,  $N_{N_2O}^c$ ,  $N_{NO}^c$  (kg N ha<sup>-1</sup> yr<sup>-1</sup>) are the emissions of ammonia, nitrous oxide and nitric oxide, respectively.

Typically, the sum of external N inputs is larger than the sum of harvested N and ammonia (and  $N_2O$  and NO) emissions. As in (Velthof et al., 2009), this N surplus is partitioned in the model between leaching of nitrate below the root zone and denitrification to dinitrogen.

#### Phosphorus, potassium, and pesticide inputs 2.3

The use of pesticides and phosphorus and potassium fertilizer is calculated as the harvested (or grazed) yield of the crop multiplied by a crop-specific usage factor in g input per kg of yield.

#### 2.4 Greenhouse gas emissions from on-farm energy use, rice, and drained organic soils

This section covers  $CO<sub>2</sub>$  emissions from energy use, methane emissions from rice cultivation, and  $CO<sub>2</sub>$  and nitrous oxide emissions from drained organic soils. For nitrous oxide emissions from mineral soils, see 2.2.6, and for methane emission from manure excreted at grazing, see 3.3.3.

### *2.4.1 CO₂ emissions from on-farm energy use in crop production*

In ClimAg, CO<sub>2</sub> emissions from on-farm diesel and electricity use for crop production include:

- Land preparation (leveling, plowing, tilling etc.), sowing and planting
- Fertilizer and pesticide application
- Manure application
- Irrigation
- Pruning of tree crops
- Harvesting and transportation to storage
- Post-harvest drying before storage

For land preparation, diesel use is calculated as a crop-specific fuel use rate per physical area. For crops on arable land, the use of diesel for tillage and sowing is calculated as a fuel use factor in *liters per ha* multiplied by the area tilled and sowed. In the case of tree crops, the energy used for establishing the plantation is apportioned over its estimated lifetime.

For the application of fertilizer and pesticides, ClimAg calculates diesel use as a fuel use rate per number of applications. Diesel use for manure application is calculated as a fuel use factor in *liter per Mg* manure multiplied by the weight of manure applied.

Energy use for irrigation includes both diesel and electricity and is calculated as a crop-specific diesel and/or electricity use rate per irrigated area.

The use of diesel for harvest (and transport to on-farm storage) is influenced by both the area covered and the weight of the harvested plant mass. Thus, diesel use is calculated as a fuel use factor in *liter per ha* multiplied by the area harvested plus two fuel use factors in *liter per Mg* multiplied by the weight of harvested products and by-products (e.g. straw), respectively. For straw not harvested, i.e., left in field, there is additional diesel use in liter per Mg for chopping of straw.

The use of fuel oil for drying harvested cereal and other grain crops is calculated as a fuel use factor in *liter per kg of water evaporated* multiplied by the amount of water evaporated in the crop. As a default, the model uses a factor of 0.15 liter oil per kg of water, based on (Edström et al., 2005).

### *2.4.2 CO₂ emissions from energy and materials use in greenhouse crop production*

To estimate the resource use and environmental impact of crop production in greenhouses, ClimAg calculates the energy use not only for operating the greenhouse but also for producing the greenhouse structures.

Energy use for operating the greenhouse includes heating, lighting, and irrigation/miscellaneous, see Table 1. To calculate the impact of the use of protective structures, ClimAg represents six different types of materials, with the specification of stocks (in kg m<sup>-2</sup>), lifetime, and energy and  $CO<sub>2</sub>$  intensity in their production.

ClimAg also represents the use of four different types of substrates in a similar way to that of protective structures.

#### *2.4.3 Methane emissions from rice cultivation*

The calculation of methane emissions from rice cultivation follows the methodology set out by (IPCC, 2019). Methane emissions, in kg CH<sub>4</sub> per *harvested* area per year, are calculated as a climate-specific default emission rate multiplied by three different scaling factors that reflect variations in management. These scaling factors reflect differences in i) water regime *during* the cultivation cycle, ii) water regime *before* the cultivation cycle, and iii) amount and type of organic matter added (e.g. manure) or left in the field (e.g. straw).

### *2.4.4 CO₂ and nitrous oxide emissions from drained organic soils*

Emissions of CO₂ and nitrous oxide from drained organic soils are calculated as the drained land area multiplied by climate-specific emission factors in kg  $N_2O$  per ha and Mg CO<sub>2</sub> per ha, respectively.

# **2.5 Carbon stock changes**

### *2.5.1 Introduction*

Since agricultural production mainly takes place on land that supports plant growth, most agricultural land use occurs at the expense of reduced carbon stored in forests and other native, carbon-rich vegetation. Therefore, agricultural land use has an inherent climate impact in the form of reduced land carbon stocks and, hence, higher atmospheric CO₂ levels. Conceptually, this effect can be described as the "carbon opportunity cost" (COC) of land: when we use a parcel of land for agricultural production, we forego the opportunity to store carbon in the native vegetation and soils that otherwise could exist on that land. (Note, however, that irrigating dry lands may, in contrast, increase carbon storage.)

Reductions (or increases) in land carbon stocks resulting from converting natural lands to agriculture and aquaculture are one-off fluxes. For example, when forests, grasslands, and other native vegetation are cleared for agriculture or aquaculture, most of the carbon stored in the vegetation is converted to CO₂ almost instantly, mainly via burning, representing a one-off pulse emission of CO₂. In contrast, if agricultural land spared from use regains its native vegetation, reaching a steady-state carbon stock will take decades or more. Yet, despite the longer time horizon, the total carbon stock increase following restoration is still a one-off change in a carbon stock: after a certain time period, there is no additional growth in the carbon stock.

In contrast to these carbon stock changes, the use of cleared land for the production of agricultural goods can proceed, in theory, indefinitely. This distinction presents a non-trivial calculation problem in apportioning the climate impact from the one-off carbon stock change (decrease or increase) over a recurring, indefinite output of agricultural goods.

The ClimAg model includes two primary approaches for addressing this calculation problem. The first approach, here called the "expansion" metric, estimates the CO<sub>2</sub> emissions that occur because of agricultural expansion (i.e. deforestation). This one-off emission can be understood conceptually as the investment cost, in units of carbon dioxide, of creating new agricultural land. The second approach, here called the "regrowth" metric, estimates the uptake of  $CO<sub>2</sub>$  that would occur if land currently in agricultural use were spared and native vegetation allowed to regrow.

For both metrics, ClimAg calculates the difference between the plant and soil carbon stored in potential native vegetation and the carbon stored in agricultural vegetation. This difference is the foregone carbon storage due to agricultural land use and represents the amount of carbon emitted in the case of the "expansion" metric, and the amount of carbon uptake in the "regrowth" metric. For both metrics the cumulative carbon storage effect from land use is the same; in practice, the only main difference between the metrics is the dynamic of the carbon stock change, as detailed below.

# *2.5.2 The "expansion" metrics: Quantifying the COC of land as the carbon emissions from converting native vegetation into agricultural land and aquaculture ponds*

In the expansion metric, the calculation issue at hand is how to apportion the one-off  $CO<sub>2</sub>$  emission from the clearing of a parcel of land (i.e., the carbon "investment cost") over the future benefits in the form of agricultural (or aquaculture) outputs from that parcel of land. Here, ClimAg uses two different approaches:

### *A. Discounted expansion metric*

Because of the uncertainty regarding tipping points in the climate system (i.e., non-linear, irreversible responses to increased global warming), a low-risk mitigation strategy should value early emission reductions more than later ones. In ClimAg this is accounted for by applying a discount rate to future CO<sub>2</sub> emissions. For consistency, ClimAg also discounts the future production on the land.

As mentioned, in the process of agricultural expansion by the destruction of native vegetation, a major fraction of the plant matter is burnt, leading to instant emissions of carbon. However, a substantial amount of plant carbon is not completely burnt but instead decomposes exponentially at a rate that depends mainly on the climate. Hence, not all of the one-off CO<sub>2</sub> emissions pulse occurs at year 0, but instead takes place over several years. For consistency, these emissions from decay are discounted to calculate an aggregate present value (see Eq. 13). Default numbers on the fraction of plant carbon burnt are shown in Table 4.

Soil carbon stock change following natural land conversion to agriculture also occurs gradually; it may take many decades to reach a new, lower soil carbon equilibrium level. ClimAg calculates soil carbon loss as a percentage loss of native soil carbon (typically 5-20%, depending on crop and climate), which occurs over a specified period (typically 30-60 years, depending on climate). The soil carbon losses are discounted to an aggregate present value assuming a linear change in soil carbon levels (see Eq. 13).

In summary, in the discounted expansion metric, the carbon opportunity cost for product (e.g., crop) *p*,  $COC_p^{exp,dis}$  (kg CO<sub>2</sub> kg<sup>-1</sup>), equals the aggregate, time-discounted carbon lost from native vegetation on land used in the region to produce the crop, divided by the aggregate, time-discounted annual production in the region for that crop:

$$
COC_p^{exp,dis} = \frac{44}{12} \times \frac{c_p^{burnt,nat} + \int_0^T \text{dis } c_p^{unburnt,nat} \times (e^d - 1) \times e^{-(d+r)t} - c_p^{plant,prod} + \int_0^T \text{dis } \frac{c_p^{solid,rad} - c_p^{solid,pr}}{T^{solid}} \times e^{-rt}}{f_0^{full}} \quad \text{Eq. 13}
$$

where:

 $C_p^{burnt, nat}$  (Mg C ha<sup>-1</sup>) is the burned amount of native vegetation plant carbon for product *p*.

 $C_p^{unburnt, nat}$  (Mg C ha<sup>-1</sup>) is the remaining, unburnt amount of native plant carbon for product *p*.

 $C_p^{plant,prod}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon in the production system for product *p*.

 $C_p^{soil,nat}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) under native vegetation for product p.

 $C_p^{solid,prod}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) in the production system for product p.

 $T_{dis}$  (years) is the discounting period.

 $T^{soil}$  (years) is the time required for soil carbon to reach a new steady state level.

 $T_{dis}^{soil}$  (years) is the discounting period for soil carbon loss (equals  $T_{oi}^{soil}$  unless  $T_{dis}^{soil}$  >  $T_{dis}$ , then  $T_{dis}^{soil}$ is set to the value of  $T_{dis}$ ).

*r* (% per year) is the discount rate.

*d* (% per year) is the decay rate for plant matter remaining after burning.

 $Y_p$  (Mg ha<sup>-1</sup> year<sup>-1</sup>) is the annual yield of product *p* (constant value over the calculation period).

The same equation applies to grazing land; in this case, grazed intake of plant matter per hectare is equivalent to yield per hectare. For aquaculture ponds, we assume that all pre-existing plants and half of the soil carbon is lost instantly. In this case, the numerator in Eq. 13 becomes  $C_p^{plant,nat} + 0.5 \times C_p^{solid,nat}$ .

<b>Biome</b>	Fraction of plant matter burnt at deforestation (at year zero)		<b>Parameter values in Chapman-</b> <b>Richards growth function</b>	
	Of above-ground	Of entire plant including roots	k	m
Tropical moist forest	52%	43%	0,090	0.5
Tropical dry forest	52%	43%	0,080	0.5
Tropical coniferous forest	52%	43%	0,050	0.5
Temperate broadleaf forest	51%	42%	0,070	0.5
Temperate coniferous forest	51%	42%	0,065	0.5
Boreal forest & taiga	59%	52%	0,050	0.5
Tropical grass- & shrubland	75%	36%	0,085	0.5
Temperate grass- & shrubland	83%	44%	0,070	0.5
Flooded grassland	75%	36%	0,075	0.5
Montane grass- & shrubland	59%	40%	0,065	0.5
Mediterranean forest & shrub	75%	40%	0,070	0.5
Deserts	75%	20%	0.060	0.5

**Table 4** Default burning rates in the expansion COC metrics and parameters in the regrowth COC metrics. Sources: (Anderson‐Teixeira and DeLucia, 2011; Cook-Patton et al., 2020).

#### *B. Amortized expansion metric*

A crude but also more straightforward approach is to amortize the total one-off carbon emission, including all cumulative soil carbon losses, evenly over a set period of years. The amortized carbon opportunity cost for product p,  $COC_p^{exp,amor}$  (kg CO<sub>2</sub> kg<sup>-1</sup>), is calculated as:

$$
COC_p^{exp,amort} = \frac{44}{12} \times \frac{C_p^{plant,nat} + C_p^{solid,nat} - C_p^{plant,prod}}{T_{amort} \times Y_p}
$$
 Eq. 14

where:

 $C_p^{plant,nat}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon in native vegetation for product *p*.  $C_p^{sol,nat}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) under native vegetation for product p.  $C_p^{plant,prod}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon in the production system for product *p*.  $C_p^{solid,prod}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) in the production system for product p.  $T_{amort}$  (years) is the amortization period.

 $Y_n$  (Mg ha<sup>-1</sup> year<sup>-1</sup>) is the annual yield of product *p* (constant value over the calculation period).

This metric is equivalent to straight-line amortization in accounting; it is also the approach recommended for accounting for carbon stock changes in the 2019 IPCC guidelines for National Inventory Reports (IPCC, 2019). It should be noted that, although not explicit, amortization, too, implies a discounting of future costs and benefits, as the discounting metric above also does. After the amortization period, future costs and benefits are assigned zero value.

#### *2.5.3 The "regrowth" metrics: Quantifying the COC of land as the carbon uptake from regrowth of potential native vegetation.*

In the regrowth metric, the carbon opportunity cost is measured as the  $CO<sub>2</sub>$  uptake that would occur if the land was no longer used, but instead allowed to regain its native vegetation. For a parcel of land, this quantity is divided by the output from the current use of that land. As with the expansion metric, ClimAg calculates two different variants:

#### *A. Discounted regrowth metric*

As mentioned above, discounting is appropriate when valuing future emissions and the future uptake of  $CO<sub>2</sub>$ . As with the expansion metric, ClimAg discounts the  $CO<sub>2</sub>$  uptake that would occur over time through the regrowth of vegetation, and the future production that takes place through continued use of the land.

ClimAg calculates the regrowth of native vegetation using the Chapman-Richards growth function, which is widely used in forestry (Burkhart and Tomé, 2012):

$$
c(t)_{p}^{plant,nat} = C_{p}^{plant,nat} \times (1 - e^{-kt})^{\frac{1}{(1-m)}}
$$
 Eq. 15

where:

 $c(t)$ <sup>plant,nat</sup> (Mg C ha<sup>-1</sup>) is the amount of plant carbon at time *t* in potential native vegetation on land where product *p* is produced.

 $C_p^{plant,nat}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon <u>at steady state</u> in potential native vegetation on land where product p is produced (equals  $C_p^{plant,nat}$  in Eq. 14).

*t* is time in years.

 $k$  and  $m$  (dimensionless) are Chapman-Richards parameters that determine the shape of the growth curve.

Table 4 shows the default numbers on parameters  $k$  and  $m$  for different biomes, which were derived by fitting the Chapman-Richard growth function to the dataset in (Cook-Patton et al., 2020). ClimAg uses these growth curves to calculate the gain of carbon in the plant component of the regrowing vegetation, as shown in Eq. 15. The gain in plant carbon over time is discounted to an aggregate present value (see Eq. 16).

For soil carbon, carbon gains are calculated in a way that is equivalent to the losses in the expansion metric (see 2.5.2). ClimAg calculates the gain as a linear increase of soil carbon back to the native, steadystate level over a time period that varies depending on the climate. The soil carbon gains are discounted to an aggregate present value (see Eq. 16).

In summary, for the discounted regrowth metric, the carbon opportunity cost for product p,  $COC_p^{regr,dis}$  (kg  $CO<sub>2</sub>$  kg<sup>-1</sup>), equals the aggregate, time-discounted carbon gain from the regrowth of native vegetation on land used in the region to produce the crop, divided by the aggregate, time-discounted annual production in the region for that crop:

$$
COC_p^{regr,dis} = \frac{44}{12} \times \frac{\int_0^{T_{dis}} [c(t)_p^{plant,nat} - c(t-1)_p^{plant,nat}] \times e^{-rt} - c_p^{plant,prod} + \int_0^{r_{dis}^{solid} p^{out}} - c_p^{coll,prod} \times e^{-rt}}{\int_0^{T_{dis}} Y_p \times e^{-rt}}
$$
 Eq. 16

where:

 $c(t)$ <sup>plant,nat</sup> (Mg C ha<sup>-1</sup>) is the amount of plant carbon at time t in potential native vegetation on land where product  $p$  is produced (see Eq. 15).

 $C_p^{soil,nat}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) under potential native vegetation on land where product  $p$  is produced.

 $C_n^{plant,prod}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon in the production system for product p.

 $C_n^{soil,prod}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) in the production system for product p.

 $T_{dis}$  (years) is the discounting period.

 $T^{soil}$  (years) is the time required for soil carbon to reach a new steady state level.

 $T_{dis}^{soil}$  (years) is the discounting period for soil carbon gain (equals  $T^{soil}$  unless  $T_{dis}^{soil} > T_{dis}$ , then  $T_{dis}^{soil}$ is set to the value of  $T_{dis}$ ).

 $r$  (% per year) is the discount rate.

 $Y_p$  (Mg ha<sup>-1</sup> year<sup>-1</sup>) is the annual yield of product p (constant value over the calculation period).

#### B. Undiscounted regrowth metric

A more straightforward method is to calculate the cumulative, undiscounted gain in carbon on a parcel of land over a set period, and divide this quantity by the cumulative, undiscounted output from the land over this period. One benefit of this approach is that it is less sensitive to the assumed shape of the growth

curve, since only the cumulative growth matters. The formula for calculating the undiscounted regrowth carbon opportunity cost for product p,  $COC_p^{regr, undis}$  (kg CO<sub>2</sub> kg<sup>-1</sup>), can be written as:

$$
COC_p^{regr,undis} = \frac{44}{12} \times \frac{c(T_{regr})_p^{plant,nat} - c_p^{plant,prod}}{T_{regr} \times Y_p} \qquad \qquad \text{Eq. 17}
$$

where:

 $c(T_{regr})_p^{plant,nat}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon at the end of the regrowth period ( $T_{regr}$ ) in potential native vegetation on land where product *p* is produced (see Eq. 15).

 $C_p^{plant,prod}$  (Mg C ha<sup>-1</sup>) is the amount of plant carbon in the production system for product *p*.

 $C_p^{solid,nat}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) under potential native vegetation on land where product  $p$  is produced.

 $C_p^{solid,prod}$  (Mg C ha<sup>-1</sup>) is the amount of soil carbon (top 1 m) in the production system for product p.

 $T_{rear}$  (years) is the regrowth period.

 $T<sup>soil</sup>$  (years) is the time required for soil carbon to reach a new steady state level.

 $\varepsilon$  (dimensionless) is the fraction of soil carbon gain that occurs during the regrowth period (equals 1) unless  $T^{soil} > T_{regr}$ , then  $\varepsilon = \frac{T_{regr}}{T^{soil}}$ .

 $Y_p$  (Mg ha<sup>-1</sup> year<sup>-1</sup>) is the annual yield of product *p* (constant value over the calculation period).

# **3. Livestock production**

The ClimAg model represents all major food-related livestock systems; see Table 5. These livestock systems account for approximately 90% of global land use by domestic animals. For all livestock systems, conventional and organic systems are represented separately. For pigs and poultry, free-range systems are also represented. ClimAg representation of cattle systems can be used as fairly accurate proxies for equivalent buffalo systems; similarly, sheep systems can be used as proxies for goat systems. In this way, the land and climate impacts of the vast majority of global food-related livestock production is well represented by the ClimAg model system.

Livestock numbers and their feed intake are calculated endogenously in the ClimAg model, using information that is available in national statistics. In this way, feed intake and feed efficiencies (feed intake per output) can be estimated using basic information specific to a country, region, or individual farm.

First, the model calculates the required number of animals per unit of output for each of the animal categories; see Figure 2, Table 5 and 3.1.1. The main exogenous parameters used here include reproduction rates, liveweight gain rates, and milk/egg yields (see Table 6). By adjusting these herd parameters, the herd productivity can be calibrated against, e.g., country-level statistics on animal numbers and production. Next, the model uses these herd parameters to calculate the required feed energy intake for each animal category, using empirical equations (see 3.2.1). Lastly, taking the calculated energy requirements as an input, feed dry matter intake is calculated with feed baskets and feed energy values as exogenous parameters (see 3.2.2). This endogenous calculation of feed intake ensures fairly accurate feed use estimates even when feed basket data are incomplete. The benefit of this model feature applies particularly to systems with significant amounts of grazing since the grazed feed quantity is rarely known.

Table 6 lists the main exogenous parameters in crop and pasture systems. The sections that follow provide more details for how livestock systems are represented in the model.



**Figure 2** Generic structure of the representation of livestock systems in the ClimAg model.Thickness of flows are roughly proportional to rate.



**Table 5** Animal categories (i.e., cohorts) included in livestock sub-systems.

For cattle and sheep systems, most key parameters, such as productivity and feed rations, are represented separately for the two principal seasons of the year: i) the wet and/or warm season when pasture production on natural and human-made grassland is abundant and ii) the dry and/or cold season when pasture production is low or non-existing. During the dry/cold season in temperate regions, animals are typically kept in barns.

### **3.1 Animals and production**

#### *3.1.1 Animal stocks and flows*

In a herd/flock module of a livestock system, the model calculates the number of animals in each animal category (cohort; see Table 5) that are required to produce one unit of output (meat, milk, or egg). Mathematically, each animal cohort category in each sub-system is modeled as a stock-flow system. This means that the model calculates the number of animals (stock) in each cohort as a function of the entry rates and exit rates (flows) in to and out of the cohort. Parameters used for setting and/or calculating entry and exit rates include birth/hatching rates, liveweight gain rates, mortality rates, culling rates, and exit ages (e.g., weaning age, slaughter age); see Table 6.

Parameters Unit/type Comments **Animals, herd dynamics**  Mature liveweight,  $MLW$  kg Current liveweight, LW and the set of the set of the set of the Endogenous parameter. Calculated using a daily time step. Liveweight at birth/hatching kg Liveweight at first birth/egg-laying kg Liveweight at slaughter kg Reproduction rate Number born/female/year For poultry, number of eggs laid Culling rate (of mature animals) % of stock/year Mortality rate – mature  $\%$  of stock/year Mortality rate – young  $\%$  of born Specified for different growth phases Milk yield kg milk/cow or ewe/day Milk consumed by offspring % DM of feed basket Wool yield kg wool/ewe/year Egg yield kg egg/hen/year Liveweight gain, LWG Specified for different growth phases  $\frac{1}{2}$  Specified for different growth phases Age at first birth/egg-laying months/days Age at slaughter months/days Carcass yield and other animal allometrics See Table 28 for numbers **Feed, enteric methane**  Maintenance energy adjustment factor for breed,  $MF_{breed}$ % of base maintenance energy requirements See 3.2.1 for details Maintenance energy adjustment factor for lactation and animal,  $MF_{animal}$ % of base maintenance energy requirements See 3.2.1 for details Maintenance energy adjustment factor for activity MF<sub>activity</sub> % of base maintenance energy requirements See 3.2.1 for details Growth energy adjustment factor for animal type  $GF_{animal}$  See 3.2.1 for details Feed basket (% of individual feeds) % DM of each feed of all intake See 3.2.2 for details Feeding waste  $\%$  of feed offered Feed energy content of each feed MJ NE or ME per kg of feed See Table 25-26, 29-30 for numbers Protein content of each feed  $\%$  DM See Table 25-26, 29-30 for numbers Emission factor enteric methane % of feed GE intake Endogenous parameter. See 3.2.4 for details

**Table 6** Main exogenous parameters in livestock sub-systems. Most parameters are specific to each animal category in the sub-system. For cattle and sheep systems, several parameters are specified separately for wet/warm and dry/cold seasons.

#### **Feed conservation (silage/hay)**



#### **Confinements, manure**

Bedding materials used in confinement kg of manure excreted



Representation of herd/flock dynamics presumes a steady-state situation, so that that herd/flock characteristics are constant from year to year. This also means that the herd/flock structure (i.e., each cohort's share of the entire herd/flock) is represented as steady-state averages, separately for the wet/warm (grazing) and dry/cold (barn) seasons in the case of cattle and sheep.

In contrast to many other herd models, the average liveweight of a cohort is not an exogenous constant but is here an endogenous parameter, *LW*, calculated using a daily time step. This allows for calculating feed energy requirements for maintenance (see 3.2.1) according to the daily liveweight figure. This approach

provides more accurate estimate of energy requirements than using an average liveweight number for a longer time period since maintenance energy requirements are non-linear with respect to liveweight (as illustrated by the equations in 3.2.1).

## *3.1.2 Production rates of meat, milk, egg, and wool*

Production rates of meat per unit of animals in stock are calculated using the herd/flock dynamics parameters in 3.1.1, which enables calculating the exit (production) rate of slaughtered animals. Combining the production rate of slaughter animals and the allometric relationships of animal organs (lean tissue, fatty tissue, bone, non-carcass part, etc.; see Table 28) give the meat production per unit of animals in the herd/flock.

Production rates of milk, egg, and wool are stated exogenously. For cattle and sheep, milk production per unit female animal per day is set separately for each of the two seasons (wet/warm and dry/cold). Egg and wool production rates are set as production per hen/ewe per year; see Table 6.

# **3.2 Feed and methane emissions from feed digestion**

# *3.2.1 Feed requirements*

For each animal category in each system, feed energy requirements are calculated using empirical bioenergetic equations for which. Basic herd parameters, such as liveweight and liveweight gain rate, are set exogenously.

Feed energy requirements represented in ClimAg are those for maintenance, activity, growth, gestation, and milk, egg, and wool production, but exclude draft work, which is of minor importance. Energy requirements are calculated using a *daily* time step and aggregated into total annual requirements (or in the case of cattle and sheep, for each of the two seasons, i.e. wet/warm and dry/cold). For cattle and sheep, the model represents feed energy requirements as net energy systems, and for pigs and poultry, metabolizable energy systems.

### *Maintenance*

The following equations are used to calculate energy requirements for maintenance, specified by species and animal category (parameter definitions follow below the equations):





 $MF_{animal}$ : Adjustment factor for lactation and animal type

 $MF_{activity}$ : Adjustment factor for activity

The type of cattle breed is known to influence maintenance requirements. For example, for Simmental breeds, energy needs are about 20% higher than in Eq. 19, whereas for Nelore breeds they are about 10% lower (National Research Council, 1996).

For lactating cattle, the model assumes maintenance energy requirements are 20% higher, based on (National Research Council, 1996). Furthermore, for uncastrated (intact) males of both cattle and sheep, maintenance energy requirements are assumed to be 15% higher (National Research Council, 2007, 1996).

For animals kept in barns or other confinements, additional energy requirements for activity is negligible. In contrast, for grazing animals, energy requirements for activity can be substantial, posing up to a 50% increase in maintenance requirements (National Research Council, 1996). In the ClimAg model, default assumptions are a 10% increase for grazing on cropland and 15-25% increase for grazing on permanent grasslands.

#### Growth

The following equations are used for calculating the energy requirements for growth (liveweight gain), specified by species and animal category (parameter definitions follow the equations):

Cattle (National Research Council, 1996):

$$
NE_g = 22 \times \left[\frac{LW}{GF_{animal} \times MLW}\right]^{0.75} \times LWG^{1.097}
$$
 [MJ NE<sub>g</sub>/day] Eq. 24

Sheep (National Research Council, 2007):

 $NE_a = (1.155 - 0.008786 \times (MLW - LW)) \times LW^{0.75} \times LWG$ [MJ  $NE_{g}/day$ ] Eq. 25



The value of the adjustment factor  $GF_{animal}$  for cattle reflects differences in percentage fat in weight gain, which tend to be higher for females than males. In the ClimAg model, this factor is 0.8 for females, 1.0 for castrated males, and 1.2 for intact males (National Research Council, 1996).

#### *Milk, egg, and wool production*



where:

 $GE_{milk}$ : Gross energy content of milk production  $GE_{egg}$ : Gross energy content of egg production  $GE_{\text{wool}}$ : Gross energy content of wool production

#### 3.2.2 Feed use

The ClimAg model calculates feed use based on the assumption that the estimated feed energy requirements for each animal category are fully met by feed matter intake. Exogenous parameters for calculating feed matter intake are the feed energy content of individual feeds and feed baskets (the share of individual feedstuff in the ration).

#### Feed energy value and composition

Data on feed energy content and other feed characteristics are given in Table 25-26, 29-30 in the Appendices.

The metabolizable and net energy value of cattle and sheep feeds from digestible energy (DE, in MJ kg dry matter) content is calculated using the following equations:

Cattle – dairy breeds (National Research Council, 2001):



Beef cattle, sheep (National Research Council, 2007, 1996):



$$
NEg = -6.90 + 1.16 \times DE - 0.0280 \times DE2 - 0.000384 \times DE3
$$
 [MJ/kg DM] Eq. 40

For calculating the metabolizable energy value of pig feeds from digestible energy (DE, in MJ kg dry matter) content, this equation is used:



#### **Feed baskets**

Table 7 presents the feed items that can be included in the feed basket for each livestock system. Feed baskets are set exogenously as a fraction of total feed intake (in percent dry matter), separately for each animal category (Table 5). For ruminant systems, feed rations are stated separately for the wet/warm (grazing) and dry/cold (barn) seasons. Additionally, feedlot rations are specified separately for farms with finishing in separate confinements ("feedlot").



**Table 7** Feed items included in the representation of feed baskets for livestock sub-systems. Fractions of feeds within the composite categories of co-products and residues are adjustable.

<sup>&</sup>lt;sup>1</sup> Includes meals from soybeans, rapeseed, sunflower, peanut, coconut, oil palm kernel, and cottonseed  $^2$  Includes wheat gluten, maize gluten meal, starch extraction residue, meat and bone meal, fish meal, and whole co

<sup>&</sup>lt;sup>3</sup> Includes broken rice, brans of wheat, rice, millet, rye; maize hominy feed, wheat gluten feed, maize gluten feed and germ meal

<sup>4</sup> Includes whey, buttermilk, and tops and leaves of cassava, potato, yams, and sugar beet
### *Feeding waste*

Some of the feed offered to animals in confinements (and supplemental feeds in grazing paddocks) is not consumed but wasted. For cereals and other concentrate feed, the waste fraction is generally very low (1-  $2\%$ ), but for silage and hay, the waste fraction can be significant (5-10%). These waste streams are accounted for in the ClimAg model. Feed waste from confinements is assumed to be incorporated into the manure stream.

### *3.2.3 Feed production and supply*

Each livestock sub-system in the ClimAg model comprises a land module to enable representation of feed production on land in the vicinity of the animal confinement, as well as grazing on cropland or permanent grassland. Representation of plant growth, nitrogen flows, and fertilizer and energy use in the land modules is described in section 2 above.

All forage feed cultivated on cropland (i.e., grasses, legumes and grass-legume mixtures, and whole cereals), and most of the cereals used as feed are assumed to be produced on the livestock farm. All other feeds, including by-products such as cereal brans, oil meals, etc., are assumed to be purchased and transported to the farm. All upstream resource use and environmental impacts of purchased feed are tracked and added to the on-farm impacts; see Table 23.

Conservation of forage crops by ensiling (or haying) involves additional diesel use and is calculated separately. Diesel use is calculated as an additional fuel use factor in *liter per ha* multiplied by the area harvested, plus a fuel use factor in *liter per Mg* multiplied by the weight of conserved plant mass (Table 6). In addition, the model accounts for dry matter losses that occur during the conservation process, typically on the order of 5 to 15%.

### *3.2.4 Methane emissions from feed digestion*

Methane emissions from feed digestion ("enteric fermentation") in ruminants and pigs are calculated as a fraction of gross energy intake. In contrast to many other livestock models, for cattle and sheep, this fraction is not an exogenous constant, but here an endogenous parameter calculated as a function of feed quality, daily feed intake, and animal liveweight. For cattle, the model uses equations developed by (Moraes et al., 2014) based on their statistical analysis of a dataset of c. 2,600 energy balance trials. For sheep, the model uses equations developed by (Van Lingen et al., 2019), who analyzed a database containing 270 measurements.

## **3.3 Manure and methane, nitrous oxide, and ammonia emissions**

### *3.3.1 Manure production*

For each animal category, the ClimAg model calculates the quantities of manure produced as a function of the energy and protein content of the feed intake, in combination with exogenous parameters on energy and ash content of feces and urine.

For cattle, sheep and pigs, the amount of feces energy produced can, by definition, be calculated as the difference between the feed intake's gross energy and digestible energy content. Similarly, the amount of urine energy can be calculated as the difference between the digestible energy content and metabolizable energy content, minus the energy in produced methane. For poultry, the difference between gross energy and metabolizable energy gives the manure energy produced.

The corresponding amount of "volatile solids" in manure (dry matter minus ash), which determines the potential production of methane (see 3.3.3), is calculated using assumed values on gross energy and ash content per kg of manure; see Table 8.

The amount of N in manure is calculated as the difference between feed N intake and N retained in animal mass. Feed N intake,  $N_{int}^{a,s}$  (Mg N yr<sup>-1</sup>), for animal cohort *a* and season *s*, is calculated from the feed basket and the protein content (divided by 6.25) of each feed component in the basket (for protein contents, see Table 25-26, 29-30). N retained in animal mass,  $N_{ret}^{a,s}$  (Mg N yr<sup>-1</sup>), is calculated from the protein content (divided by 6.4) of liveweight gain and milk/egg/wool production (for protein contents, see Table 28).

The partition of excess N intake into feces and urine is calculated according to the relationship developed by (Scholefield et al., 1991):

$$
N_{uri}^{a,s} = (N_{int}^{a,s} - N_{ret}^{a,s}) \times (14 \times NC_{rat}^{a,s} + 0.24)
$$
 Eq. 42

$$
N_{fcc}^{a,s} = (N_{int}^{a,s} - N_{ret}^{a,s} - N_{uri}^{a,s})
$$
 Eq. 43

where  $N_{uri}^{a,s}$  (kg N yr<sup>-1</sup>) and  $N_{fcc}^{a,s}$  (kg N yr<sup>-1</sup>) are the amounts of N excreted in urine and feces, respectively, and  $NC_{rat}^{a,s}$  (dimensionless) is the N concentration (on a dry matter basis) in the feed basket.



**Table 8** Composition of manure fractions (as excreted) in livestock sub-systems. For sources, see table footnotes.

<sup>&</sup>lt;sup>1</sup> Calculated assuming an ash free gross energy content of 21.0 MJ/kg DM for cattle and pig feces (Font-Palma, 2019; Wnetrzak et al., 2015) and 19.5 for poultry manure (Quiroga et al., 2010), and 11 MJ/kg DM for urine (close to urea, 10.5 MJ/kg DM).<br>
<sup>2</sup> Based on (Hansen et al., 2008; Mathot et al., 2020, 2012; Petersen et al., 2016)<br>
<sup>3</sup> Same numb

The fraction of feces and urine excreted on pastures is assumed to equal the fraction of time spent on pastures. The fraction of time spent on pastures is estimated as the fraction of grazed feed intake of total feed intake.

Feed waste and bedding materials are added to the manure excreted in confinements. The amount of volatile solids and N in these streams are included in the calculations of methane and N emissions from manure; see 3.3.3 and 3.3.4.

### *3.3.2 Type of confinements and manure storage*

The type of animal confinement and manure storage technology influence the emission rates of methane, nitrous oxide, and ammonia from manure. In the ClimAg model, confinements and storage are represented by ten different manure management system types; see Table 6.

Emissions of methane, nitrous oxide, and ammonia are calculated separately for each of these confinement and storage technologies to reflect how their different inherent conditions influence emission rates.

## *3.3.3 Methane emissions and manure decay*

Soon after excretion, the volatile solids (VS) in manure are subjected to decomposition by microbes. The ClimAg model represents this decomposition by setting exogenously a fraction of VS decay,  $DF^m$ , that occurs in the confinement and subsequent storage (if any).

Decomposition of manure leads to production of methane and  $CO<sub>2</sub>$ , as well as conversion of organic N into ammonium (inorganic N), of which some is converted to ammonia and emitted (see next section). Rate of methane production is calculated as a function of the excreted quantity of VS, multiplied by an animal- and feed-specific factor that reflects the maximum potential methane production per unit of VS (denoted B) and a climate- and management-specific methane conversion factor (often denoted MCF) that reflects to what extent the maximum methane production is realized. The part of the fraction of VS decay that does not cause methane production is assumed to be lost as CO₂.

These calculations are done separately for the emissions that occur in animal confinements and emissions that occur during subsequent manure storage, if any. Apart from the methane generated from the manure itself, calculations are also done of methane produced from substrates added to the manure stream in confinements, mainly bedding materials and feeding waste.

### *Confinement*

Methane emissions in confinements increase with longer manure retention times. Therefore, methane emissions are calculated as a function of retention time.

Emissions of methane,  $C_{CH4,con}^{a}$  (kg yr<sup>-1</sup>), for animal cohort *a*, is calculated according to the following equation:

$$
C_{CH_4,con}^a = 0.67 \times \left(VS_{con,fcc}^a \times B_{fcc}^a + VS_{con,uri}^a \times B_{uri}^a + VS_{bed}^a \times B_{bed}^a + VS_{wst}^a \times B_{wst}^a\right) \times
$$
  
\n
$$
EF_{CH4,con}^m \times \tau_{con}^m
$$
 Eq. 44

where:

 $VS_{con, fec}^a$  and  $VS_{con,uri}^a$  (kg DM yr<sup>-1</sup>) is the volatile solids content, in feces and urine, respectively, excreted in confinement by animal *a*.

 $VS_{bed}^a$  and  $VS_{wst}^a$  (kg DM yr<sup>-1</sup>) is the volatile solids content of used bedding materials (on-farm supplied or purchased) and feed waste (uneaten feed in confinement) for animal *a*.

 $B^a$  (m<sup>3</sup> CH<sub>4</sub> (kg VS)<sup>-1</sup>) is the maximum CH<sub>4</sub> production potential per unit of volatile solids in the substrate.

 $EF_{CH_{\alpha}con}^{m}$  (dimensionless) is the *daily* methane production in confinement as a fraction of the maximum CH4 production potential of the substrate in manure system *m*.

 $\tau_{con}^{m}$  (days) is the retention time in the confinement of substrate (excreted manure, bedding material, and feed waste) for manure system *m*.

Calculation of the output of substrate with methane-production potential that leaves the confinement (i..e, which equals input to storage), considers the losses of VS that occur in the confinement due to decay. Hence, the "maximum CH4 production potential" in the substrate leaving the confinement as output,  $maxCH4<sup>a</sup><sub>man.stl</sub>$  (kg yr<sup>-1</sup>)t, for animal cohort *a* is given by:

$$
maxCH_{4man,con} = 0.67 \times (VS_{st,rec}^a \times B_{st,rec}^a + VS_{st,uri}^a \times B_{st,uri}^a + VS_{bed}^a \times B_{bed}^a +
$$
  
\n
$$
VS_{wst}^a \times B_{wst}^a \times (1 - EF_{CH_+con}^m - DF_{con}^a)
$$
 Eq. 45

where:

 $DF_{con}^a$  (dimensionless) is the fraction VS being decomposed in the confinement, of the amount VS excreted in confinement by animal *a*.

### *Storage*

To calculate methane emissions during manure storage, the ClimAg model uses different approaches for solid and liquid manure types.

For solid manure types (feces, deep litter, poultry manure, etc.), methane production rates are generally low, and therefore a simple approach is used. Emissions of methane from manure storage,  $C_{CH4,sto}^{a}$  (kg yr  $\alpha$ <sup>1</sup>), are calculated according to:

$$
C_{CH_4,Sto}^a = maxCH_{4man,con}^a \times EF_{CH_4,Sto}^{m,s}
$$
 Eq. 46

where:

 $EF_{CH4, sto}^{m,s}$  (dimensionless) is the total production of methane during storage as a fraction of the maximum CH4 production potential of the substrate in manure system *m, s* (*s* as in solid manure). The emission factor here equals the commonly used MCF parameter and is an exogenous fixed number for the annual amount of manure entering storage, although specific to climate zone and manure type.

For liquid manure types (slurry, urine, etc.), methane production rates can be very high, and a more detailed approach is used to reflect the large regional variation in methane emissions due to climatic

differences. Here, the ClimAg model calculates the methane emission factor  $EF_{CH4, st0}^{m,l}$  during the storage period using *monthly* average temperatures as input, in contrast to most other models which use annual average temperature. Here, methane production calculations are based on the predictive model presented in (IPCC, 2019), which is itself based on a model developed by (Mangino et al., 2001).

Since methane production is non-linearly related to temperature, calculating methane emissions using average temperatures over a long time period (e.g., a year) is likely to underestimate emissions. In general, modeled estimates based on shorter time steps will provide more accurate emission estimates. This is particularly true in cool regions where most annual methane production occurs during a few warm months when temperatures exceed 15 °C.

In addition to using monthly temperature data, the calculation also factors in the timing and frequency of removal (emptying) of manure from storage for application on land. The resulting emission factor  $EF_{CH4,sto}^{m,l}$  is used to calculate annual methane emissions in the same way as in Eq. 46:

$$
C_{CH4,sto}^a = maxCH_{4man,con}^{a} \times EF_{CH4,sto}^{m,l}
$$
 Eq. 47

*Grazing* 

Methane emissions from manure excreted on pastures,  $C_{CH4,araz}^{a}$  (kg yr<sup>-1</sup>), are generally very small, and are calculated in a simple way:

$$
C_{CH_4, graz}^a = 0.67 \times \left(VS_{graz, fec}^a \times B_{fcc}^a + VS_{graz, uri}^a \times B_{uri}^a\right) \times EF_{CH_4, graz}^m
$$
 Eq. 48

where:

 $VS_{con, fec}^a$  and  $VS_{con,uri}^a$  (kg DM yr<sup>-1</sup>) is the volatile solids content, in feces and urine, respectively, excreted at grazing by animal *a*.

 $B^a$  (m<sup>3</sup> CH<sub>4</sub> (kg VS)<sup>-1</sup>) is the maximum CH<sub>4</sub> production potential per unit of volatile solids in substrate.

 $EF_{CH4, graz}^{m}$  (dimensionless) is the total production of methane as a fraction of maximum CH<sub>4</sub> production potential of substrate.

### *3.3.4 Nitrogen emissions and output*

In the ClimAg model, descriptions of inputs, losses, and outputs of N in animal confinements and manure storage facilities are made on a mass-balance basis, with explicit distinction between inorganic and organic N. Input flows represented include feces, urine, bedding material and feed waste, separately by animal cohort.

#### *Ammonia and nitrous oxide emissions in confinement*

Most of the N in excreted manure is in inorganic form, i.e., ammonium (see Table 8). Decomposition of manure leads to conversion of the organic N in the manure into ammonium, increasing the supply of inorganic N.

Following common practice, in the ClimAg model, manure ammonia emissions are calculated as a fraction (emission factor) of the amount of ammonium in the manure. In the confinement phase, the emission factor is applied to the amount of ammonium *when excreted*. The emissions of ammonia from confinement,  $N_{NH_3,con}^a$  (kg N yr<sup>-1</sup>), for animal cohort *a*, is calculated according to:

$$
N_{NH_3,con}^a = (N_{st. fec,ino}^a + N_{st.uri,ino}^a) \times EF_{NH_3,con}^m
$$
 Eq. 49

where:

 $N_{st. fec,ino}^a$  and  $N_{st.uri,ino}^a$  (kg N yr<sup>-1</sup>) is the inorganic N content, in feces and urine, respectively, excreted in confinement by animal *a*.

 $EF_{NH3,con}^{m}$  (dimensionless) is the fraction lost through ammonia volatilization in confinement for manure system *m*.

Emissions of nitrous oxide from the confinement,  $N_{N20,con}^a$  (kg N yr<sup>-1</sup>), are calculated as an emission factor of the *total* amount of N entering the confinement:

$$
N_{N_2O,con}^a = (N_{st, fec,ino}^a + N_{st.uri,ino}^a + N_{st, fec,org}^a + N_{st.uri,org}^a + N_{bed}^a + N_{wst}^a) \times EF_{N_2O,con}^m
$$
 Eq. 50

where:

 $EF_{N20,con}^{m}$  (dimensionless) is the fraction emitted as nitrous oxide in confinement for manure system m.

### *Output of nitrogen from confinement*

The calculation of the output of N from the confinement considers the losses as ammonia and nitrous oxide, and the increase of ammonium and the equivalent decrease in organic N because of decay of volatile solids. It also considers N losses through runoff. Hence, output from the confinement of organic-N,  $N_{man, stl, org}^{a}$  (kg N yr<sup>-1</sup>), and inorganic N,  $N_{man, stl,ino}^{a}$  (kg N yr<sup>-1</sup>), which equal input to storage (or to field, if daily spread), for animal cohort *a* is given by:

$$
N_{man,con,org}^{a} = (N_{st, fec,org}^{a} + N_{st,uri,org}^{a} + N_{bed}^{a} + N_{wst}^{a}) \times (1 - DF_{con}^{m}) - N_{run,org}^{a} \qquad \text{Eq. 51}
$$
\n
$$
N_{man,con,ino}^{a} = N_{st, fec,ino}^{a} + N_{st,uri,ino}^{a} + (N_{st, fec,org}^{a} + N_{st,uri,org}^{a} + N_{bed}^{a} + N_{wst}^{a}) \times DF_{con}^{m} - N_{NH_3,con}^{a} - N_{N_2O,con}^{a} - N_{run,ino}^{a} \qquad \text{Eq. 52}
$$

where:

 $DF_{con}^{m}$  (dimensionless) is the fraction of volatile solids decomposed in the confinement for manure system *m*.

 $N_{run,con,org}^a$  and  $N_{run,con,ino}^a$  are organic-N and inorganic-N content in surface runoff from the confinement

Note that runoff may in some systems be collected in a settling basin and, subsequently, a holding pond, and may be recycled to the field; here, however, it is treated as a lost flow.

#### *Ammonia and nitrous oxide emissions from storage*

Emissions of ammonia from storage,  $N_{NH_3, st0}^a$  (kg N yr<sup>-1</sup>), are calculated as an emission factor of the amount of ammonium (inorganic nitrogen) entering storage:

$$
N_{NH_3,sto}^a = N_{man,con,ino}^a \times EF_{NH_3,sto}^m
$$
 Eq. 53

where  $EF_{NH3,sto}^{m}$  (dimensionless) is the fraction lost through ammonia volatilization during storage for manure system *m*.

Emissions of nitrous oxide from storage,  $N_{N20,sto}^a$  (kg N yr<sup>-1</sup>), is calculated as an emission factor of the amount of *total* N entering storage:

$$
N_{N20,sto}^{a} = (N_{man,con,ino}^{a} + N_{man,con,org}^{a}) \times EF_{N20,sto}^{m}
$$
 Eq. 54

where  $EF_{N20,sto}^{m}$  (dimensionless) is the fraction emitted as nitrous oxide during storage for manure system *m*.

#### *Output of nitrogen from storage*

The calculation of the output of N from storage considers the losses as ammonia and nitrous oxide, and the increase of ammonium and the equivalent decrease in organic nitrogen because of the decay of volatile solids. It also considers N losses through runoff. Output from storage of organic-N,  $N_{man, sto, org}^{l}$  (kg N yr <sup>1</sup>), and inorganic N,  $N_{man, sto, ino}^{l}$  (kg N yr<sup>-1</sup>), for application to field (or other use) is the sum of manure storage output for all animal categories,  $a = 1, 2, ..., n$ , in livestock system *l*, and is calculated as:

$$
N_{man,sto,org}^l = \sum_{a=1}^n \left( N_{man,sto,org}^a \times (1 - DF_{sto}^m) \right)
$$
 Eq. 55

$$
N_{man,sto,ino}^l = \sum_{a=1}^n (N_{man,sto,ino}^a + N_{man,sto,org}^a \times DF_{sto}^m - N_{NH_3,sto}^a - N_{N_20,sto}^a)
$$
 Eq. 56

where:

 $DF_{sto}^{m}$  (dimensionless) is the fraction of volatile solids decomposed in during storage for manure system *m*.

 $N_{run.sto,org}^a$  and  $N_{run.sto,ino}^a$  are organic-N and inorganic-N content in surface runoff from storage

Note that N emissions occurring after application of manure on cropland, and after excretion on pastures, are described in 2.2.6.

#### **3.4 CO₂ emissions from on-farm energy use**

For livestock farming operations, in addition to those for feed production, ClimAg calculates CO<sub>2</sub> emissions from energy use separately for three categories:

- Fuel oil for heating
- Electricity for milking

 Fuel oil and/or electricity for all other purposes (feeding, ventilation, lighting, manure management, etc.)

Energy use and emissions from heating and general purposes are calculated by assuming systems-specific energy use per animal unit and time spent in confinement. Annual energy use is calculated by multiplying these factors by the percentage time of the year spent in confinement. In this way, the model factors in the differences in energy use due to varying extent of grazing in ruminant systems.

## **4. Aquaculture production, capture fisheries**

Compared to livestock systems, the potential climate and environmental impacts of aquaculture systems are typically less varied. In the ClimAg model, therefore, the representation of aquaculture systems is simpler compared to that of livestock systems. Table 9 presents the main exogenous parameters included in the representation of aquaculture systems.

### **4.1 Feed and land use in aquaculture**

In aquaculture, feed use efficiency is typically quantified according to the "economic feed conversion ratio" (eFCR), which quantifies total feed input per total net output (actual harvest) of product. The ratio factors in losses of the product by death, escapes, etc., and that of non-ingested feed. Because of relatively small variation in feed requirements across aquaculture systems, improvements in model accuracy are less dependent on detailed estimates of feed energy requirements, in contrast to the modeling of livestock systems. Instead, feed use in aquaculture is represented simply in ClimAg using exogenous values for eFCR.



**Table 9** Main exogenous parameters in aquaculture sub-systems. Sub-systems are listed in Table 18.

Several common species in aquaculture can feed on organic matter naturally present in the water body, such as plankton and detritus. Some filter-feeding species, such as certain carp (e.g., silver carp) and mollusks, feed exclusively on naturally occurring food, and their production uses no external feed. The ClimAg model represents the use of in-situ feed by a parameter that states the fraction of external feed in each sub-system.

Feed baskets are set exogenously as a fraction of total *external* feed use. Table 10 presents the possible feed items available for the feed basket in the aquaculture system. All external feed is assumed to be transported to aquaculture facilities from crop farms and/or compound feed plants. Upstream resource use and environmental impacts of external feed that occur in crop production and processing are accounted for and added to the on-site impacts; see Table 23.

**Table 10** Feed items included in the representation of feed baskets for aquaculture sub-systems. Fractions of feeds within the composite categories are adjustable.

<b>Items</b>
<b>Unprocessed crop products</b>
Wheat grains, Maize grains, Soybean seeds, Faba beans, Pea seeds, Cassava (dried)
Crop starch concentrates
Wheat flour, Wheat starch, Maize starch, Broken rice
Oil Vegetable oil <sup>1</sup> , Fish oil <sup>2</sup>
Crop protein concentrates
Wheat bran, Maize hominy feed, Rice bran, Wheat gluten meal, Maize gluten meal, Oil meals <sup>3</sup>
Animal protein concentrates Meat and bone meal <sup>4</sup> , Fish meal <sup>5</sup>
Pigments etc Pigments, Amino acids, Minerals/vitamins

Aquaculture production of crustaceans and freshwater fish mainly occurs in artificial ponds, created at the expense of native vegetation or other land uses. In ClimAg, land use is defined by an exogenous parameter that sets the land requirement per annual output of product. Carbon storage changes due to this land use are calculated in the same way as for agricultural land use (see 2.5).

### **4.2 Methane and nitrous oxide emissions from aquaculture**

Large input of feed to aquaculture ponds, in combination with poor aeration of the water mass, stimulates substantial methane production. Methane production is represented by an exogenous parameter that sets the annual methane emission per hectare of water area.

Because of the large input of N in feed to aquaculture ponds, nitrous oxide  $(N_2O)$  production in the water mass is larger than what it would be without the feed input. In ClimAg, nitrous oxide emissions are calculated as an emission factor multiplied by the amount of feed N input to the water mass that is not retained in animal mass, i.e., feed N excreted in feces and feed not ingested. N content retained in animal mass is calculated as the protein content of the aquacultural output (see Table 28) divided by 6.4.

<sup>&</sup>lt;sup>1</sup> Includes oils from all vegetable oil sub-systems included in ClimAg (see Table 19)<br>
<sup>2</sup> Includes oils from all fish sub-systems included in ClimAg (see Table 20)<br>
<sup>3</sup> Includes meals from soybeans, rapeseed, sunflower,

ClimAg model description

### **4.3 CO₂ emissions from energy use in aquaculture and capture fisheries**

For aquaculture operations, ClimAg calculates CO<sub>2</sub> emissions from energy use separately for:

- Gas, fuel oil, and electricity for production of compound feed
- Diesel and electricity for running the aquaculture facility

For capture fisheries, ClimAg includes CO<sub>2</sub> emissions from the fuel consumed by fishing vessels.

## **5. Processing of crop, livestock, and aquaculture and fisheries products**

### **5.1 Food products: Crop products and plant-based meat and dairy substitutes**

Processing of crop products mainly involves separating plant materials into more homogenous fractions that have a relatively high concentration of either starch, oil, sugar, or protein. This processing is hereafter referred to as "primary" processing. Some of the outputs from primary crop processing are consumed as food (e.g., vegetable oils and white rice), and some are used as feedstock in further processing (e.g., composite products), hereafter referred to as "secondary" processing.

In ClimAg, both primary and secondary processing are described on a mass and energy balance basis, with separate balances for nitrogen (protein). In primary processing, the yield of the main product, as well as that of significant co-products, are represented. Energy use in each process is represented, with separate calculations for process steps with significant energy use, such as drying; see Table 11. Upstream resource use and environmental impacts associated with the production of the feedstocks are accounted for and added to the on-site (i.e., the processing plant) impacts. see Table 23.



**Table 11** Main exogenous parameters in primary processing of crop products into food-type items. Sub-systems are listed in Table 19.

As to secondary processing, the ClimAg model also represents the production of plant-based meat and dairy substitutes, as these products generally have a lower climate cost compared to animal meat and dairy products.

Plant-based meat substitutes are currently marketed in many different forms. Products designed to closely resemble real animal meat are typically made from a combination of protein concentrates (and/or isolates) and vegetable oils, together with additives and other minor ingredients. Among the most used plant

protein sources are soybeans and peas. As a fat source, any vegetable oil may be used, except in certain products, such as patties, for which coconut fat is preferred for its high melting point.

For plant-based meat substitutes, ClimAg represents four distinct, but generalized ingredient configurations for plant-based meat products; see Table 21. These configurations use either soybean or peas as a plant protein source, either at a low or high fat content. In the high-fat configurations, coconut fat is used.

For plant-based dairy product substitutes, ClimAg represents the most common types of milk substitutes (soy, oat, almond, rice), and three variants of plant-based butter substitutes based either on soy oil, palm oil or coconut oil, in addition to rapeseed and sunflower oil which are included in all three variants; see Table 21. For cheese and cream substitutes, only one ingredient configuration is included, reflecting the smaller variability within the ingredient composition of currently marketed products.

Exogenous parameters taken to model plant-based meat and dairy sub-processes are feedstock inputs (kg feedstock per kg output) and energy use per output. As in the case of primary processing, upstream resource use and environmental impacts associated with the production of the feedstocks are accounted for and added to the on-site (the processing plant) impacts; see Table 23.

### **5.2 Food products: Dairy, meat, and fish & shellfish**

Primary processing of slaughtered animals and fish/shellfish involves the cutting of body parts to separate non-food (i.e., hides, guts) from food parts and further cutting and/or grinding of food parts to obtain specific meat and fish/shellfish products. Primary processing of whole milk generally represents more diverse processes than that of slaughtered animals. Basic processes involve production of items with a lower or higher milk fat concentration than whole milk, removal of the milk carbohydrate fraction (cheese production), and drying into milk powder products.



**Table 12** Main exogenous parameters in primary processing of dairy, meat and fish/shellfish products. Sub-systems are listed in Table 20.

As for crops, the yield of the main product, as well as that of significant co-products, are represented in the primary processing of livestock and fish/shellfish. Also, energy use in each process is represented,

with separate calculations for process steps with significant energy use, such as rendering and drying; see Table 12. As in the case of crop processing, upstream resource use and environmental impacts associated with the production of the feedstocks are accounted for and added to the on-site (the processing plant) impacts; see Table 23.

## **5.3 Materials products: Cotton**

Some major global agricultural crops are produced mainly for materials functions. These include seed cotton, linseed, and rubber trees. In ClimAg, representation of seed cotton is included.

The model also includes the processing of seed cotton into cotton lint (the main product, used for textile purposes), and various co-products; see Table 22. Exogenous parameters for seed cotton processing are analogous to those for primary processing of crops to food; see Table 11. As in the case of crop processing, upstream resource use and environmental impacts associated with the production of the feedstocks are accounted for and added to the on-site (the processing plant) impacts; see Table 23.

## 5.4 **Energy products: Liquid fuels, gas**

A significant fraction of the global production of agricultural crops is used to produce liquid fuels, destined mainly for the road transportation sector. ClimAg represents nine different types of biodiesel and bioethanol; see Table 22. Exogenous parameters taken in modeling the production of these liquid fuels are analogous to those for primary processing of crops into food; see Table 11. As in the case of crop processing, upstream resource use and environmental impacts associated with the production of the feedstocks are accounted for and added to the on-site (the processing plant) impacts; see Table 23.

Some manure and other biomass streams are currently diverted into anaerobic reactors, which are designed to realize to the greatest extend possible the methane production potential of the inherent substrates. ClimAg includes representation of reactors that use cattle or pig slurry as substrates, with whole-cereal silages and food waste as complementary substrates. Exogenous parameters for reactors are analogous to those for methane production from manure; see 3.3.3.

## **6. Production of fossil-based fuels, electricity, fertilizers, and pesticides**

The ClimAg model represents six different fuels made from fossil carbon feedstocks: coal, oil, gas, diesel, gasoline, and kerosene (jet fuel). These fuels are characterized by their energy and carbon content per unit weight and volume. In addition to fuel-CO<sub>2</sub> released at burning, ClimAg represents emissions associated with the extraction of feedstocks (mainly methane leaks) and processing into ready-to-use fuels (mainly refinery emissions).

For electricity, ClimAg includes one average CO2 intensity for all electricity use in all sectors. This number represents the average on-site and upstream emissions associated with electricity production in a region.

ClimAg represents seven different fertilizers, mainly of single-nutrient type, and one pesticide type; see Table 13. The representation of phosphorus, potassium, and pesticides in ClimAg is simple due to their relatively low energy intensity or relatively low consumption (pesticides).



**Table 13** Fertilizers and pesticides included in ClimAg and feedstocks and parameters represented in their production.

## **7. Trade and transportation**

### **7.1 Trade balances and resource/environmental costs of imports**

In any global-scale, multi-regional application of ClimAg, trade between regions is represented for all major items that are traded over longer distances. In the case of crops, this includes, for example, most cereals and other dry crops, but excludes bulky crops, such as forages (silage, etc.) and sugar crops.

Upstream resource use and environmental impacts associated with imports to a region are calculated as the weighted average of the resource use and environmental impacts per kg for the exported quantities from all exporting regions. Energy use and emissions associated with the importation transport (see 7.2) are added to the upstream resource use and emissions. Hence, the resource use and emissions per kg of imported items is the weighted average of upstream resource use and emissions plus that of the importation itself.

### **7.2 Energy use and CO₂ emissions from freight transport**

### *7.2.1 Transport nodes and cargo characteristics*

In the ClimAg model, energy use and emissions from freight transport are included for all routes that significantly add to the total environmental impact:

- Transport of crop products from the farm or greenhouse to either: i) food primary processing plant (to make, e.g., flour or oil); ii) food stores for direct consumption (e.g., vegetables and fruits); iii) livestock and aquaculture farms for use as feed; or iv) other processing plants (e.g., biofuels)
- Transport of whole animals and whole milk to processing plants, and whole eggs to food stores
- Transport of processed food from processing plants to food stores for consumption, or secondary processing plants
- Importation of products to a region

Each route is described in terms of distances, divided into distances for "long" distribution and "short" distribution. "Short" distribution refers to shorter legs from or to the point of departure or arrival of the cargo as part of longer routes that use several modes of transport. In ClimAg, short distribution is done by trucks; for long distribution, there are several possible transport modes (see 7.2.2).

In addition to distances, the cargo is described in terms of its pallet density and whether the cargo needs to be chilled during transport. Pallet density is the weight of the cargo per volume required in its packaged form and determines whether weight or volume is the limiting factor for the mode of transport (in maritime shipping, the inverted concept, the "stowage" factor, is used). Chilled transport creates an additional energy requirement on top of that for the locomotion.

### *7.2.2 Modes of transport*

Table 14 describes the freight transport options included in ClimAg. Ground transport options do not include rail transport because of its small importance for the transport of agricultural goods. Furthermore, only fuel-propelled options are included; no electric options are included. Electric trucks are increasingly being deployed in several regions, but still comprise a small percentage of overall truck fleets.





DWT: deadweight tonnage; TEU: twenty-foot equivalent (container equivalent) <sup>1</sup> Based on NTM Calc at https://www.transportmeasures.org/en/

2 Author estimates based on various industry data. 3 Based on (Swahn, 2008)

## **8. Food end-use and food waste**

In ClimAg, consumption is referred to as "end-use," to distinguish the use of items for consumption from that as feedstock. Food end-use in ClimAg is represented using about 130 items, of which 45 are items from livestock and fish & seafood; see Table 15.

Food end-use in ClimAg is represented as "apparent" consumption, which is the amount of food delivered to the food retail sector. This quantity is estimated on an annual basis using statistics on production.

ClimAg also represents the actual intake of food, that is, the amount of food ingested. Based on detailed descriptions of the chemical composition of food items (see Table 25-29), ClimAg calculates the daily intake per capita of protein, fat, carbohydrates, alcohol, and crude fiber.

Food waste is calculated as the difference between reported apparent consumption in statistics and estimated actual intake per capita, and includes waste in retail, households, restaurants and other food outlets. ClimAg distinguishes between edible and inedible/unpreferred items in food waste. Inedible/unpreferred items include bones, egg shell, hulls, peelings and similar, and is calculated based on allometric data for each item. Edible waste is calculated as a fraction of apparent consumption and is calculated separately for all items in the food basket, to be able to reflect the varying levels of waste for different types of food.

Based on detailed data on upstream resource use and environmental impacts for each food item (Table 23), ClimAg calculates the per-capita resource use and environmental impacts from food end-use. However, ClimAg does not include energy use in retail, for food storage or food preparation. Additionally, the resource use and environmental impacts of food packaging are not included.



**Table 15** Food items included in the representation of food consumption.



### **9. References**

AHDB, 2020a. Beef yield guide. Agriculture and Horticulture Development Board.

- AHDB, 2020b. Lamb yield guide. Agriculture and Horticulture Development Board.
- AHDB, 2020c. Pork yield guide. Agriculture and Horticulture Development Board.
- Aletor, V.A., Hamid, I.I., Nieß, E., Pfeffer, E., 2000. Low-protein amino acid-supplemented diets in broiler chickens: effects on performance, carcass characteristics, whole-body composition and efficiencies of nutrient utilisation. Journal of the Science of Food and Agriculture 80, 547–554.
- Anderson‐Teixeira, K.J., DeLucia, E.H., 2011. The greenhouse gas value of ecosystems. Global change biology 17, 425–438.
- Ashworth, A. j., Chastain, J. p., Moore Jr., P. a., 2020. Nutrient Characteristics of Poultry Manure and Litter, in: Animal Manure. John Wiley & Sons, Ltd, pp. 63–87. https://doi.org/10.2134/asaspecpub67.c5
- Atti, N., Hamouda, M.B., 2004. Relationships among carcass composition and tail measurements in fat-tailed Barbarine sheep. Small Ruminant Research 53, 151–155.
- Balthazar, C., Pimentel, T., Ferrão, L., Almada, C., Santillo, A., Albenzio, M., Mollakhalili, N., Mortazavian, A., Nascimento, J.S., Silva, M., 2017. Sheep milk: Physicochemical characteristics and relevance for functional food development. Comprehensive reviews in food science and food safety 16, 247–262.
- Bartzas, G., Vamvuka, D., Komnitsas, K., 2017. Comparative life cycle assessment of pistachio, almond and apple production. Inf Process Agric 4: 188–198.
- Bhagya, H.P., Maheswarappa, H.P., Surekha, Bhat, R., 2017. Carbon sequestration potential in coconut-based cropping systems. Ind. Jour. of Hort. 74, 1. https://doi.org/10.5958/0974-0112.2017.00004.4
- Bowen, P.A., Zebarth, B.J., Toivonen, P.M.A., 1999. Dynamics of nitrogen and dry-matter partitioning and accumulation in broccoli (Brassica oleracea var. italica) in relation to extractable soil inorganic nitrogen. Can. J. Plant Sci. 79, 277–286. https://doi.org/10.4141/P98-056
- Brito de Figueirêdo, M.C., Potting, J., Lopes Serrano, L.A., Bezerra, M.A., da Silva Barros, V., Gondim, R.S., Nemecek, T., 2016. Environmental assessment of tropical perennial crops: the case of the Brazilian cashew. Journal of Cleaner Production 112, 131–140. https://doi.org/10.1016/j.jclepro.2015.05.134
- Burkhart, H.E., Tomé, M., 2012. Modeling forest trees and stands. Springer Science & Business Media.
- Chase, L.D., Henson, I.E., 2010. A detailed greenhouse gas budget for palm oil production. International Journal of Agricultural Sustainability 8, 199–214.
- Cook-Patton, S.C., Leavitt, S.M., Gibbs, D., Harris, N.L., Lister, K., Anderson-Teixeira, K.J., Briggs, R.D., Chazdon, R.L., Crowther, T.W., Ellis, P.W., 2020. Mapping carbon accumulation potential from global natural forest regrowth. Nature 585, 545–550.
- Dao, A., Bationo, B.A., Traoré, S., Bognounou, F., Thiombiano, A., 2021. Using allometric models to estimate aboveground biomass and predict carbon stocks of mango (Mangifera indica L.) parklands in the Sudanian zone of Burkina Faso. Environmental Challenges 3, 100051. https://doi.org/10.1016/j.envc.2021.100051
- Daouda, B.O., Aliou, S., Léonard, A.E., Yasmine, A.J.F., Vincent, A., Irénikatché, A.P.B., Nestor, A., 2017. Assessment of organic carbon stock in cashew plantations in Benin. International Journal of Agriculture and Environmental Research 03.
- De Lange, C., Morel, P., Birkett, S., 2003. Modeling chemical and physical body composition of the growing pig. Journal of Animal Science 81, E159–E165.
- Edman, F., Ahlgren, S., Landquist, B., 2023. Kött-och slaktutbyte–data och metoder vid beräkningar av miljöpåverkan.
- Edström, M., Pettersson, O., Nilsson, L., Hörndahl, T., 2005. Jordbrukssektorns energianvändning.
- Ehrenbergerová, L., Cienciala, E., Kučera, A., Guy, L., Habrová, H., 2016. Carbon stock in agroforestry coffee plantations with different shade trees in Villa Rica, Peru. Agroforest Syst 90, 433–445. https://doi.org/10.1007/s10457-015-9865-z
- Everts, H., Dekker, R., 1995. Effect of protein supply during pregnancy and lactation on body composition of sows during three reproductive cycles. Livestock Production Science 43, 137–147.
- Font-Palma, C., 2019. Methods for the Treatment of Cattle Manure—A Review. C 5, 27. https://doi.org/10.3390/c5020027
- Ganeshamurthy, A., 2023. Annual carbon capture potential of banana in India. BIOTROPIA 30, 374–383. https://doi.org/10.11598/btb.2023.30.3.2005
- García‐Condado, S., López‐Lozano, R., Panarello, L., Cerrani, I., Nisini, L., Zucchini, A., Van der Velde, M., Baruth, B., 2019. Assessing lignocellulosic biomass production from crop residues in the European Union: Modelling, analysis of the current scenario and drivers of interannual variability. GCB Bioenergy 11, 809–831.
- Gephart, J.A., Henriksson, P.J.G., Parker, R.W.R., Shepon, A., Gorospe, K.D., Bergman, K., Eshel, G., Golden, C.D., Halpern, B.S., Hornborg, S., Jonell, M., Metian, M., Mifflin, K., Newton, R., Tyedmers, P., Zhang, W., Ziegler, F., Troell, M., 2021. Environmental performance of blue foods. Nature 597, 360–365. https://doi.org/10.1038/s41586-021-03889-2
- Głowacz-Różyńska, A., Tynek, M., Malinowska-Pańczyk, E., Martysiak-Żurowska, D., Pawłowicz, R., Kołodziejska, I., 2016. Comparison of oil yield and quality obtained by different extraction procedures from salmon (Salmo salar) processing byproducts. European Journal of Lipid Science and Technology 118, 1759– 1767. https://doi.org/10.1002/ejlt.201500269
- Gonçalves, A.L., Kernaghan, J.R., 2014. Banana production methods.
- Hansen, M.N., Sommer, S.G., Hutchings, N.J., Sørensen, P., 2008. Emissionsfaktorer til beregning af ammoniakfordampning ved lagring og udbringning af husdyrgødning [Emission factors for calculation of ammonia volatilization by storage and application of animal manure]. DJF Husdyrbrug 45.
- Henson, I.E., Dolmat, M.T., 2003. Physiological analysis of an oil palm density trial on a peat soil. Journal of Oil Palm Research 15.
- Hernández‐Montes, E., Escalona, J.M., Tomás, M., Martorell, S., Bota, J., Tortosa, I., Medrano, H., 2022. Carbon balance in grapevines: effect of environment, cultivar and phenology on carbon gain, losses and allocation. Aust J Grape and Wine Res 28, 534–544. https://doi.org/10.1111/ajgw.12557
- Høgh-Jensen, H., Loges, R., Jørgensen, F.V., Vinther, F.P., Jensen, E.S., 2004. An empirical model for quantification of symbiotic nitrogen fixation in grass-clover mixtures. Agricultural Systems 82, 181–194. https://doi.org/10.1016/j.agsy.2003.12.003
- Hussein, E., Suliman, G., Alowaimer, A., Ahmed, S., Abd El-Hack, M., Taha, A., Swelum, A., 2020. Growth, carcass characteristics, and meat quality of broilers fed a low-energy diet supplemented with a multienzyme preparation. Poultry science 99, 1988–1994.
- Iglesias, D.J., Quiñones, A., Font, A., Martínez-Alcántara, B., Forner-Giner, M.Á., Legaz, F., Primo-Millo, E., 2013. Carbon balance of citrus plantations in Eastern Spain. Agriculture, Ecosystems & Environment 171, 103–111. https://doi.org/10.1016/j.agee.2013.03.015
- INRA, n.d. Feedipedia [WWW Document]. URL https://www.feedipedia.org/
- IPCC, 2019. 2019 Refinement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories.
- Johansen, U., Nistad, A.A., Ziegler, F., Mehta, S., Wocken, Y., Hognes, E.S., 2022. Greenhouse gas emissions of Norwegian salmon products.
- Kage, H., Stützel, H., 1999. A simple empirical model for predicting development and dry matter partitioning in cauliflower (*Brassica oleracea* L. *botrytis*). Scientia Horticulturae 80, 19–38. https://doi.org/10.1016/S0304- 4238(98)00226-X
- Kapur, R., Duttamajumder, S., Srivastava, B., Madhok, H., Kumar, R., 2013. Harvest index and the components of biological yield in sugarcane. Indian Journal of Genetics and Plant Breeding (The) 73, 386. https://doi.org/10.5958/j.0975-6906.73.4.058
- Karlsson Potter, H., Kätterer, T., Lang, R., 2023. Climate impact of liming arable soil effect on N2O emissions in a life cycle perspective (Mistra Food Futures Report No. 12), Mistra Food Futures Report. Swedish University of Agricultural Sciences, Uppsala.
- Kendall, A., Marvinney, E., Brodt, S., Zhu, W., 2015. Life Cycle–based Assessment of Energy Use and Greenhouse Gas Emissions in Almond Production, Part I: Analytical Framework and Baseline Results. J of Industrial Ecology 19, 1008–1018. https://doi.org/10.1111/jiec.12332
- Kho, L.K., Jepsen, M.R., 2015. Carbon stock of oil palm plantations and tropical forests in M alaysia: A review. Singap J Trop Geogr 36, 249–266. https://doi.org/10.1111/sjtg.12100
- Kiyanzad, 2005. Comparison of Carcass Composition of Iranian Fat-tailed Sheep.
- Kyriazakis, I., Whittemore, C.T. (Eds.), 2006. Whittemore's Science and Practice of Pig Production, 1st ed. Wiley. https://doi.org/10.1002/9780470995624
- Lam, W.Y., van Zelm, R., Benítez-López, A., Kulak, M., Sim, S., King, J.H., Huijbregts, M.A., 2018. Variability of greenhouse gas footprints of field tomatoes grown for processing: interyear and intercountry assessment. Environmental science & technology 52, 135–144.
- Larbier, M., Leclercq, B., 1994. Nutrition and Feeding of Poultry. Nottingham University Press.
- Lasco, R.D., 2002. Forest carbon budgets in Southeast Asia following harvesting and land cover change. SCIENCE IN CHINA SERIES C LIFE SCIENCES-ENGLISH EDITION- 45, 55–64.
- Mahgoub, O., Lu, C.D., Early, R.J., 2000. Effects of dietary energy density on feed intake, body weight gain and carcass chemical composition of Omani growing lambs. Small Ruminant Research 37, 35–42. https://doi.org/10.1016/S0921-4488(99)00132-7
- Mangino, J., Bartram, D., Brazy, A., 2001. Development of a methane conversion factor to estimate emissions from animal waste lagoons. Presented at the US EPA's 17th Annual Emission Inventory Conference, Atlanta GA, USA.
- Massuquetto, A., Panisson, J.C., Marx, F.O., Surek, D., Krabbe, E.L., Maiorka, A., 2019. Effect of pelleting and different feeding programs on growth performance, carcass yield, and nutrient digestibility in broiler chickens. Poultry Science 98, 5497–5503.
- Mathot, M., Decruyenaere, V., Stilmant, D., Lambert, R., 2012. Effect of cattle diet and manure storage conditions on carbon dioxide, methane and nitrous oxide emissions from tie-stall barns and stored solid manure. Agriculture, Ecosystems & Environment 148, 134–144. https://doi.org/10.1016/j.agee.2011.11.012
- Mathot, M., Lambert, R., Stilmant, D., Decruyenaere, V., 2020. Carbon, nitrogen, phosphorus and potassium flows and losses from solid and semi-solid manures produced by beef cattle in deep litter barns and tied stalls. Agricultural Systems 178, 102735. https://doi.org/10.1016/j.agsy.2019.102735
- Mavromichalis, I., Emmert, J.L., Aoyagi, S., Baker, D.H., 2000. Chemical Composition of Whole Body, Tissues, and Organs of Young Chickens (*Gallus domesticus*). Journal of Food Composition and Analysis 13, 799–807. https://doi.org/10.1006/jfca.2000.0934
- Moraes, L.E., Strathe, A.B., Fadel, J.G., Casper, D.P., Kebreab, E., 2014. Prediction of enteric methane emissions from cattle. Global change biology 20, 2140–2148.
- Morandé, J.A., Stockert, C.M., Liles, G.C., Williams, J.N., Smart, D.R., Viers, J.H., 2017. From berries to blocks: carbon stock quantification of a California vineyard. Carbon Balance and Management 12, 5. https://doi.org/10.1186/s13021-017-0071-3
- Nahm, K.H., 2003. Evaluation of the nitrogen content in poultry manure. World's Poultry Science Journal 59, 77–88. https://doi.org/10.1079/WPS20030004
- Naik, S.K., Sarkar, P.K., Das, B., Singh, A.K., Bhatt, B.P., 2019. Biomass production and carbon stocks estimate in mango orchards of hot and sub-humid climate in eastern region, India. Carbon Management 10, 477–487. https://doi.org/10.1080/17583004.2019.1642043
- Nair, P.K.R., Mohan Kumar, B., Naresh Kumar, S., 2018. Climate Change, Carbon Sequestration, and Coconut-Based Ecosystems, in: Krishnakumar, V., Thampan, P.K., Nair, M.A. (Eds.), The Coconut Palm (Cocos Nucifera L.) - Research and Development Perspectives. Springer, Singapore, pp. 779–799. https://doi.org/10.1007/978-981-13-2754-4\_16
- National Academies of Sciences, 2021. Nutrient Requirements of Dairy Cattle: Eighth Revised Edition. National Academies Press, Washington, D.C. https://doi.org/10.17226/25806
- National Academies of Sciences, 2016. Nutrient Requirements of Beef Cattle: Eighth Revised Edition. The National Academies Press, Washington, DC. https://doi.org/10.17226/19014
- National Research Council, 2007. Nutrient Requirements of Small Ruminants: Sheep, Goats, Cervids, and New World Camelids. The National Academies Press, Washington, DC. https://doi.org/10.17226/11654
- National Research Council, 2001. Nutrient Requirements of Dairy Cattle: Seventh Revised Edition. The National Academies Press, Washington, DC. https://doi.org/10.17226/9825
- National Research Council, 1996. Nutrient Requirements of Beef Cattle, Seventh Rev.
- N'Gbala, F.N., Guéi, A.M., Tondoh, J.E., 2017. Carbon stocks in selected tree plantations, as compared with semideciduous forests in centre-west Côte d'Ivoire. Agriculture, Ecosystems & Environment 239, 30–37. https://doi.org/10.1016/j.agee.2017.01.015
- Nordgarden, U., Hemre, G.-I., Hansen, T., 2002. Growth and body composition of Atlantic salmon (*Salmo salar* L.) parr and smolt fed diets varying in protein and lipid contents. Aquaculture 207, 65–78. https://doi.org/10.1016/S0044-8486(01)00750-5
- Ortiz-Ceballos, G.C., Vargas-Mendoza, M., Ortiz-Ceballos, A.I., Mendoza Briseño, M., Ortiz-Hernández, G., 2020. Aboveground carbon storage in coffee agroecosystems: The case of the central region of the state of Veracruz in Mexico. Agronomy 10, 382.
- Ortiz-Ulloa, J.A., Abril-González, M.F., Pelaez-Samaniego, M.R., Zalamea-Piedra, T.S., 2021. Biomass yield and carbon abatement potential of banana crops (Musa spp.) in Ecuador. Environ Sci Pollut Res 28, 18741–18753. https://doi.org/10.1007/s11356-020-09755-4
- Ovi, F., Hauck, R., Grueber, J., Mussini, F., Pacheco, W., 2021. Effects of prepelleting whole corn inclusion on feed particle size, pellet quality, growth performance, carcass yield, and digestive organ development and intestinal microbiome of broilers between 14 and 42 d of age. Journal of Applied Poultry Research 30, 100113.
- Palosuo, T., Heikkinen, J., Regina, K., 2015. Method for estimating soil carbon stock changes in Finnish mineral cropland and grassland soils. Carbon Management 6, 207–220. https://doi.org/10.1080/17583004.2015.1131383
- Petersen, S.O., Olsen, A.B., Elsgaard, L., Triolo, J.M., Sommer, S.G., 2016. Estimation of Methane Emissions from Slurry Pits below Pig and Cattle Confinements. PLoS ONE 11, e0160968. https://doi.org/10.1371/journal.pone.0160968
- Proietti, S., Sdringola, P., Desideri, U., Zepparelli, F., Brunori, A., Ilarioni, L., Nasini, L., Regni, L., Proietti, P., 2014. Carbon footprint of an olive tree grove. Applied Energy 127, 115–124. https://doi.org/10.1016/j.apenergy.2014.04.019
- Prussi, M., Yugo, M., De, P.L., Padella, M., Edwards, R., 2020. JEC Well-To-Wheels report v5. https://doi.org/10.2760/100379
- Public Health England, 2021. Composition of foods integrated dataset (CoFID) [WWW Document]. GOV.UK. URL https://www.gov.uk/government/publications/composition-of-foods-integrated-dataset-cofid (accessed 10.30.24).
- Quiroga, G., Castrillón, L., Fernández-Nava, Y., Marañón, E., 2010. Physico-chemical analysis and calorific values of poultry manure. Waste Management 30, 880–884. https://doi.org/10.1016/j.wasman.2009.12.016
- Reid, J.B., English, J.M., 2000. Potential Yield in Carrots ( Daucus carota L.): Theory, Test, and an Application. Annals of Botany 85, 593–605. https://doi.org/10.1006/anbo.2000.1108
- Richards, N.K., 1992. Cashew tree nutrition related to biomass accumulation, nutrient composition and nutient cycling in sandy red earths of Northern Territory, Australia. Scientia Horticulturae 52, 124–142. https://doi.org/10.1016/0304-4238(92)90015-5
- Rodrigues, M.Â., Ferreira, I.Q., Claro, A.M., Arrobas, M., 2012. Fertilizer recommendations for olive based upon nutrients removed in crop and pruning. Scientia Horticulturae 142, 205–211. https://doi.org/10.1016/j.scienta.2012.05.024
- Rodríguez, A., Peña-Fleitas, M.T., Gallardo, M., de Souza, R., Padilla, F.M., Thompson, R.B., 2020. Sweet pepper and nitrogen supply in greenhouse production: Critical nitrogen curve, agronomic responses and risk of nitrogen loss. European Journal of Agronomy 117, 126046. https://doi.org/10.1016/j.eja.2020.126046
- Sahoo, U.K., Nath, A.J., Lalnunpuii, K., 2021. Biomass estimation models, biomass storage and ecosystem carbon stock in sweet orange orchards: Implications for land use management. Acta Ecologica Sinica 41, 57–63. https://doi.org/10.1016/j.chnaes.2020.12.003
- Sanchez, M., González, J.L., 2005. The fertilizer value of pig slurry. I. Values depending on the type of operation. Bioresource Technology 96, 1117–1123. https://doi.org/10.1016/j.biortech.2004.10.002
- Sauvant, D., 2004. Tables of composition and nutritional value of feed materials. INRA.
- Schmidt, J., 2007. Life cycle assessment of rapeseed oil and palm oil: Ph. D. thesis, Part 3: Life cycle inventory of rapeseed oil and palm oil.
- Scholefield, D., Lockyer, D.R., Whitehead, D.C., Tyson, K.C., 1991. A model to predict transformations and losses of nitrogen in UK pastures grazed by beef cattle. Plant and Soil 132, 165–177. https://doi.org/10.1007/BF00010397
- Sen, A., Santra, A., Karim, S., 2004. Carcass yield, composition and meat quality attributes of sheep and goat under semiarid conditions. Meat science 66, 757–763.
- Somarriba, E., Cerda, R., Orozco, L., Cifuentes, M., Dávila, H., Espin, T., Mavisoy, H., Ávila, G., Alvarado, E., Poveda, V., Astorga, C., Say, E., Deheuvels, O., 2013. Carbon stocks and cocoa yields in agroforestry systems of Central America. Agriculture, Ecosystems & Environment 173, 46–57. https://doi.org/10.1016/j.agee.2013.04.013

Strid, I., Röös, E., Tidåker, P., 2014. Förluster av svenskt nötkött inom primärproduktion och slakt.

- Swahn, M., 2008. Additional CO2e-factors in goods transport. Network for Transport and Environment.
- Tacon, A.G., Metian, M., 2013. Fish matters: importance of aquatic foods in human nutrition and global food supply. Reviews in fisheries Science 21, 22–38.
- Taghizadeh-Toosi, A., Christensen, B.T., Hutchings, N.J., Vejlin, J., Kätterer, T., Glendining, M., Olesen, J.E., 2014. C-TOOL: A simple model for simulating whole-profile carbon storage in temperate agricultural soils. Ecological Modelling 292, 11–25. https://doi.org/10.1016/j.ecolmodel.2014.08.016
- U.S Department of Agriculture, n.d. FoodData Central [WWW Document]. URL https://fdc.nal.usda.gov/
- Van Lingen, H.J., Niu, M., Kebreab, E., Valadares Filho, S.C., Rooke, J.A., Duthie, C.-A., Schwarm, A., Kreuzer, M., Hynd, P.I., Caetano, M., 2019. Prediction of enteric methane production, yield and intensity of beef cattle using an intercontinental database. Agriculture, Ecosystems & Environment 283, 106575.
- Velthof, G., Oudendag, D., Witzke, H., Asman, W., Klimont, Z., Oenema, O., 2009. Integrated assessment of nitrogen losses from agriculture in EU‐27 using MITERRA‐EUROPE. Journal of environmental quality 38, 402–417.
- Victor, A.D., Valery, N.N., Boris, N., Aimé, V.B.T., Louis, Z., 2021. Carbon storage in cashew plantations in Central Africa: case of Cameroon. Carbon Management 12, 25–35. https://doi.org/10.1080/17583004.2020.1858682
- Villaverde, C., Baucells, M.D., Cortinas, L., Hervera, M., Barroeta, A.C., 2005. Chemical composition and energy content of chickens in response to different levels of dietary polyunsaturated fatty acids. Archives of Animal Nutrition 59, 281–292. https://doi.org/10.1080/17450390500217082
- Vu, V.T.K., Prapaspongsa, T., Poulsen, H.D., Jørgensen, H., 2009. Prediction of manure nitrogen and carbon output from grower-finisher pigs. Animal Feed Science and Technology 151, 97–110. https://doi.org/10.1016/j.anifeedsci.2008.10.008
- Whittemore, C., Yang, H., 1989. Physical and chemical composition of the body of breeding sows with differing body subcutaneous fat depth at parturition, differing nutrition during lactation and differing litter size. Animal Science 48, 203–212.
- Wirsenius, S., 2000. Human Use of Land and Organic Materials: Modeling the Turnover of Biomass in the Global Production System. PhD thesis. Chalmers University of Technology and Goteborg University.
- Wnetrzak, R., Hayes, D.J.M., Jensen, L.S., Leahy, J.J., Kwapinski, W., 2015. Determination of the Higher Heating Value of Pig Manure. Waste Biomass Valor 6, 327–333. https://doi.org/10.1007/s12649-015-9350-y
- Zanotelli, D., Montagnani, L., Manca, G., Scandellari, F., Tagliavini, M., 2015. Net ecosystem carbon balance of an apple orchard. European Journal of Agronomy 63, 97–104. https://doi.org/10.1016/j.eja.2014.12.002
- Zhao, M.Q., Li, M., Shi, Y.F., 2014. Carbon Storage and Carbon Dioxide Sequestration of Banana Plants at Different Growth Stages. Advanced Materials Research 1010–1012, 662–665. https://doi.org/10.4028/www.scientific.net/AMR.1010-1012.662

# **Appendices**



**Table 16** Crop/pasture sub-systems included in ClimAg.





**Table 17** Livestock sub-systems included in ClimAg.

### **Table 18** Aquaculture/fisheries sub-systems included in ClimAg.





**Table 19** Sub-systems of crop processing for food-type items included in ClimAg. Feedstock stated when consisting of other than whole crop product.



### **Table 20** Sub-systems of livestock and fish/shellfish processing included in ClimAg



Reduction fish Fish oil, fish meal

**Table 21** Sub-systems of manufacturing of composite food items included in ClimAg.











**Table 23** Resource use, emissions and foregone carbon stocks represented for each category of sub-systems.

Category	Resource use, emissions, and carbon stock changes represented						
	On-site (within sub-system) impact parameters included			Inputs to sub-system	Up-stream impacts included for inputs		
	<b>Resource use</b> including inputs	<b>Emissions</b>	<b>Carbon</b> stock changes	Item	<b>Resource use</b>	<b>Emissions</b>	<b>Carbon stock</b> changes
<b>Production of</b> farmed fish/shellfish (aquaculture)	Land (ponds only)	$N_2O$ water mass	Foregone plant and soil C stocks (ponds only)	Gas, diesel, electricity		$CO2$ fuels and	
	Gas Diesel Electricity	CH <sub>4</sub> water mass (ponds only)				electricity production	
		$CO2$ fuel use					
	Aquaculture feed			Feed (see Error! Reference source not found.)	Cropland, permanent grassland, new fixed nitrogen <sup>1</sup> , energy	$N_2O^2$ , $CH_4^3$ , $CO_2^4$ Ammonia, nitrate	Foregone plant and soil C stocks
Primary processing of crops, and livestock and aquaculture outputs	Gas Fuel oil Electricity	CO <sub>2</sub> fuel use		Gas, diesel, electricity		CO <sub>2</sub> fuels and electricity production	
	Feedstock (crop products, animals)			Feedstock (crop products, animals)	Cropland, permanent grassland, new fixed nitrogen <sup>1</sup> , energy	$N_2O^2$ , $CH_4^3$ , $CO_2^4$ Ammonia, nitrate	Foregone plant and soil C stocks
Secondary processing of outputs from primary processing (bio- diesel, plant- based, etc.)	Gas Fuel oil Electricity	$CO2$ fuel use		Gas, diesel, electricity		$CO2$ fuels and electricity production	
	Feedstock			Feedstock	Cropland, permanent grassland, new fixed nitrogen <sup>1</sup> , energy	$N_2O^2$ , $CH_4^3$ , $CO_2^4$ Ammonia, nitrate	Foregone plant and soil C stocks
Consumption of food, energy, materials	Food, biofuels, cotton			Food, biofuels, cotton	Cropland, permanent grassland, new fixed nitrogen <sup>1</sup> , energy	$N_2O^2$ , $CH_4^3$ , $CO_2^4$ Ammonia, nitrate	Foregone plant and soil C stocks

<sup>&</sup>lt;sup>1</sup> Includes nitrogen fertilizer and biologically fixed nitrogen.

4 Includes all potential upstream sources of CO2 (organic soils, energy use). Specified separately for organic soils, fertilizer/pesticides production, crop and livestock farm energy use, primary and secondary processing energy use, and transportation energy use.

 $^2$  Includes all potential upstream sources of N2O (fertilizer production, soils, manure, aquaculture ponds, "indirect" N2O). Specified separately for fertilizer production, crop farms, and livestock farms.

<sup>&</sup>lt;sup>3</sup> Includes all potential upstream sources of CH4 (flooded rice, animals, manure, aquaculture ponds, crop processing). Specified separately for flooded rice, livestock/aquaculture farms, and crop processing.



**Table 24** Default values on key parameters that represent plant growth in tree and bush crops. For sources, see table footnotes.

<sup>&</sup>lt;sup>1</sup> Based on (Chase and Henson, 2010; Henson and Dolmat, 2003; Kho and Jepsen, 2015; Schmidt, 2007)

 $2$  Based on (Bhagya et al., 2017; Lasco, 2002; Nair et al., 2018)

<sup>&</sup>lt;sup>3</sup> Based on (Proietti et al., 2014; Rodrigues et al., 2012)

<sup>4</sup> Based on (Brito de Figueirêdo et al., 2016; Daouda et al., 2017; Richards, 1992; Victor et al., 2021)

<sup>5</sup> Based on (Bartzas et al., 2017; Kendall et al., 2015)

<sup>6</sup> Based on (Hernández‐Montes et al., 2022; Morandé et al., 2017)

 $\frac{7}{7}$  Based on (Dao et al., 2021; Naik et al., 2019)

<sup>8</sup> Based on (Ganeshamurthy, 2023; Gonçalves and Kernaghan, 2014; Ortiz-Ulloa et al., 2021; Zhao et al., 2014)

<sup>&</sup>lt;sup>9</sup> Based on (Zanotelli et al., 2015)

 $10$  Based on (Iglesias et al., 2013; Sahoo et al., 2021)

<sup>&</sup>lt;sup>11</sup> Based on (Ehrenbergerová et al., 2016; N'Gbala et al., 2017; Ortiz-Ceballos et al., 2020; Somarriba et al., 2013)

**Table 25** Allometrics, composition and energy value of crop components. Allometrics and composition as % of dry matter unless otherwise stated. Energy values in MJ/kg dry matter. For tree and bush crops, percentage of product refers to the annual production of the production divided by the annual turnover of all above ground mass, and percentage of roots refers to the annual turnover of root mass divided by the annual turnover of all plant mass. Numbers shown on crop product as percentage of above-ground mass are example data only, valid for East Asia. For sources, see table footnotes.










	<b>ALLOMETRICS1</b>		<b>COMPOSITION AND ENERGY VALUE<sup>2</sup></b>											
	<b>Of</b> above- ground	Of whole plant		DM (at DM (at Protein harvest) storage)		Lipid	Carbo- hydrate	<b>Dietary</b> fiber	<b>NDF</b>	GE- <b>HHV</b>	<b>GE-LHV Human</b>	ME	Rumi- nant DE	Pig DE Chicken <b>ME</b>
Cassava		12%			5.0%									
White potato		12%			8.8%									
Sweet potato		12%			8.8%									
Yam		12%			8.8%									
Sugar cane		8%			6.0%									
Sugar beet		12%			6.0%									
Tomato		5,0%			6.0%									
Okra		$10\%$			6.0%									
Pea (green)		16%			5.0%									
Cabbage		$6.0\%$			10%									
Cucumber		6.0%			6.0%									
Pepper (capsicum)		$6.0\%$			6.0%									
Eggplant		6.0%			6.0%									
Cauliflower & broccoli		5.0%			13%									
Other vegetables		5.0%			6.0%									
Onion		5.0%			$6.0\%$									
Carrot		5.0%			6.0%									
Grape		$8.0\%$			3.1%									
Mango		11%			3.1%									
Plantain		6.8%			3.1%									
Banana		5.8%			3.1%									
Apple		8.1%			3.1%									
Orange		7.9%			3.1%									
Other fruits		7.5%			3.1%									
Cocoa		14%			3.1%									
Coffee		12%			3.1%									
Tea		13%			3.1%									
Cotton		17%			5.0%									
Grass-legume on cropland <sup>3</sup>		45%												
Grasses					7.0%									
Legumes					14%									
Perm. and semi-perm. grass		$50\%$ <sup>4</sup>			5.0%									

<sup>&</sup>lt;sup>1</sup> Based on (Bowen et al., 1999; García-Condado et al., 2019; IPCC, 2019; Kage and Stützel, 1999; Kapur et al., 2013; Lam et al., 2018; Palosuo et al., 2015; Reid and English, 2000; Rodríguez et al., 2020; Taghizadeh-Toosi et al., 2014)

<sup>&</sup>lt;sup>2</sup> Food item composition based on (Public Health England, 2021; U.S Department of Agriculture, n.d.); feed item composition based and (INRA, n.d.; National Academies of Sciences, 2021, 2016; Sauvant, 2004; Wirsenius, 2000)

<sup>3</sup> Additional composition data is available in Wirsenius et al, The full climate cost of agriculture and aquaculture including foregone land carbon storage, *In preparation*

<sup>4</sup> Amount of root mass is assumed to be twice that of above-ground mass, and that the annual turnover of root mass is 50%



Table 26 Composition and energy value of products and by-products from crop processing into food products. Composition as % of dry matter unless otherwise stated. Energy numbers in MJ/kg dry matter. For sources, see table footnotes.



<sup>1</sup> Food item composition based on (Public Health England, 2021; U.S Department of Agriculture, n.d.); feed item composition based and (INRA, n.d.; National Academies of Sciences, 2021, 2016; Sauvant, 2004; Wirsenius, 2000)

Table 27 Composition and energy value of plant-based meat and dairy substitutes. Composition as % of dry matter unless otherwise stated. Energy numbers in MJ/kg dry matter unless otherwise stated. Based on back-of-package information for a large set of plant-based products currently on the market.





**Table 28** Allometrics, composition and energy value of livestock and fish body components. Allometrics as % of fresh weight. Composition as % of dry weight. Numbers shown on allometrics are example data only, valid for East Asia. For sources, see table footnotes.



					ALLOMETRICS OF WHOLE BODY AND CARCASS <sup>1</sup>			<b>COMPOSITION AND ENERGY VALUE<sup>2</sup></b>							
	whole body	Carcass/ Carcass - fillet - of of <i>empty</i> body	Lean tissue (of carcass (of carcass or cut)	Fatty tissue or cut)	<b>Bone</b> (of carcass or cut)	<b>Skin</b> $($ of carcass)	cass parts (of whole (of empty body)	body)	Non-car- Digesta Meat cut $($ of carcass)	DM $(% )^{2}(x)$ fresh)	<b>Total</b> protein	Edible protein $(%$ fresh weight)	<b>Total</b> lipid	Total ash Human	ME (MJ/ kg fresh weight)
<b>Chickens</b>															
Breast, skin on			97%	3.0%	$0\%$				28%	27.9%	73%	20%	25%	$2.0\%$	6.0
Thigh			68%	12%	20%				27%	37.1%	51%	19%	38%	$11\%$	8.1
Drumstick			60%	10%	30%				15%	38.0%	51%	19%	34%	15%	7.4
Wing			48%	7.0%	45%				10%	39.8%	49%	20%	30%	21%	6.6
<b>BODY TISSUES</b>															
Cattle															
Lean tissue										26.0%	78.1%	20.3%	17.7%	$4.2\%$	5.2
Fatty tissue										76.0%	11.6%	8.8%	88.0%	0.4%	26.2
Bone tissue										69.0%	30.4%		26.1%	43.5%	
Non-carcass parts										45.0%	35.0%		55.0%	10.0%	
<b>Sheep</b>															
Lean tissue										26.0%	78.1%	20.3%	17.7%	4.2%	5.2
Fatty tissue										76.0%	11.6%	8.8%	88.0%	0.4%	26.2
Bone tissue										69.0%	30.4%		26.1%	43.5%	
Non-carcass parts										54.0%	38.0%		54.0%	$8.0\%$	
Wool										98%	93%		$6.0\%$	1.0%	
Pigs															
Lean tissue										28.5%	75.0%	21.4%	20.0%	5.0%	5.7
Fatty tissue										78.9%	8.6%	6.8%	90.5%	0.9%	27.6
Bone tissue										70.0%	30.4%		26.1%	43.5%	
Skin tissue										25.0%	77.5%		20.0%	2.5%	
Other non-carcass										45.0%	57.5%		40.0%	2.5%	
<b>Chickens</b>															
Lean tissue										26.0%	80.0%	20.8%	18.0%	2.0%	5.3
Fatty tissue										80.0%	8.0%	6.4%	90.0%	2.0%	27.7
Bone tissue										48.5%	42.0%		22.0%	36.0%	
Non-carcass parts										30.0%	73.0%		22.0%	5.0%	

 $^1$  Based on (AHDB, 2020a, 2020b, 2020c; Atti and Hamouda, 2004; De Lange et al., 2003; Edman et al., 2023; Everts and Dekker, 1995; Gephart et al., 2021; Hussein et al., 2020; Johansen et al., 2020; Johansen et al., 2022; Kyriazakis and Whittemore, 2006; Massuquetto et al., 2019; Ovi et al., 2021; Sen et al., 2004; Strid et al., 2014; Whittemore and Yang, 1989; Wirsenius, 2000)

<sup>&</sup>lt;sup>2</sup> Based on (Aletor et al., 2000; Głowacz-Różyńska et al., 2016; Kiyanzad, 2005; Mahgoub et al., 2000; Mavromichalis et al., 2000; Nordgarden et al., 2002; Tacon and Metian, 2013; Villaverde et al., 2005; Wirsenius, 2000)

Table 29 Composition and energy value of products and co-products from processing of livestock and fish. Composition as % of dry matter unless otherwise stated. Energy numbers in MJ/kg dry matter unless otherwise stated. For sources, see table footnotes.





<sup>1</sup> Based on (Public Health England, 2021; U.S Department of Agriculture, n.d.)

<sup>2</sup> Whole milk composition based on (Balthazar et al., 2017); other on (Public Health England, 2021; U.S Department of Agriculture, n.d.)

<sup>3</sup> Calculated from data in Table 28

<sup>4</sup> Based on (Tacon and Metian, 2013)

 $^5$  Based on (INRA, n.d.; National Academies of Sciences, 2021, 2016; Sauvant, 2004; Wirsenius, 2000)

6 Offal composition based on (Public Health England, 2021; U.S Department of Agriculture, n.d.); other based on (INRA, n.d.; National Academies of Sciences, 2021, 2016; Sauvant, 2004; Wirsenius, 2000)



**Table 30** Composition and energy value of materials and energy products and related by-products. Composition as % of dry matter unless otherwise stated. Energy numbers in MJ/kg dry matter unless otherwise stated. For sources, see table footnotes.

<sup>1</sup> Based on (INRA, n.d.; National Academies of Sciences, 2021, 2016; Sauvant, 2004; Wirsenius, 2000)

<sup>2</sup> Based on (INRA, n.d.; National Academies of Sciences, 2021, 2016; Prussi et al., 2020; Sauvant, 2004; Wirsenius, 2000)

<sup>&</sup>lt;sup>3</sup> Based on (Prussi et al., 2020)



**Figure 3** Schematic of the representation of nitrogen flows in crop and livestock modules in the ClimAg model